

Seasonal release from competition explains partial migration in European moose

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Abstract:	<p>Partial migration, whereby a proportion of a population migrates between distinct seasonal ranges, is common throughout the animal kingdom. However, studies linking existing theoretical models of migration probability, with empirical data are lacking. The competitive release hypothesis for partial migration predicts that due to density-dependent habitat selection, the proportion of migrants increases as the relative quality and size of the seasonal range increases, but decreases with increasing migration cost and population density. To test this prediction, we developed a quantitative framework to predict the proportion of migrants, using empirical data from 545 individually GPS-marked moose (<i>Alces alces</i>) from across Fennoscandia, spanning latitudes of 56° to 68°N. Moose contracted their ranges to common and spatially limited winter areas (typically at lower elevation), but expanded them during summer due to an increase in suitable habitat (at highland ranges). As predicted from our model, a better and larger highland range relative to the lowland range corresponded to a higher proportion of migrants in an area. Quantitative predictions coupling the balance of habitat availability of seasonal ranges with the probability of migrating in a large herbivore is a necessary step towards an enhanced understanding of the mechanisms underlying migration at the population level.</p>

1 **Introduction**

2 Migration is a behavioral adaptation to seasonal changes in the geographic distribution of
3 suitable habitat either through genotypic (Boyle and Conway 2007, Holt and Fryxell 2011) or
4 phenotypic changes in habitat selection (Van Moorter et al. 2020). Partial migration, whereby
5 a fraction of a population migrates, is widespread across the animal kingdom (Lack 1943,
6 Chapman et al. 2011a and 2011b, Berg et al. 2019). Several mechanisms have been proposed
7 for the persistence of different movement strategies within a single population. Chapman et
8 al. (2011b) reviewed eight such hypotheses to explain partial migration: arrival time,
9 competitive release, fasting endurance, predation vulnerability, sexual conflict, social fence,
10 thermal tolerance, and trophic polymorphism (see also Berg et al. 2019 for a discussion
11 focused on ungulates). Underlying all these hypotheses is either individual variation in
12 tolerance of environmental constraints or some form of inter-individual competition resulting
13 in density-dependent habitat selection. In the absence of individual variation in tolerance to
14 adverse conditions, theoretical models suggest that density-dependent habitat selection is
15 necessary to explain partial migration (e.g. Kaitala et al. 1993, Taylor and Norris 2007, Holt
16 and Fryxell 2011, Fryxell and Holt 2013, Van Moorter et al., 2020).

17 Partial migration can exist in populations where residents and migrants either share a
18 common summer ground and separate during winter (i.e. non-breeding partial migration or
19 shared breeding) or share a common wintering area and separate during summer (i.e.
20 breeding partial migration or shared non-breeding). Although, both forms of partial migration
21 have been described in a range of species (Griswold et al. 2010, Chapman et al. 2011b), it is
22 shared non-breeding that is most common among ungulates at northern latitudes (Mysterud et
23 al. 2011, Berg et al. 2019). Ungulates are often forced from areas with high snow depth
24 during winter (e.g. Mysterud, Bjørnsen and Østbye, 1997) due to reduced access to food
25 (Schwab, Pitt, and Schwab 1987) and increased cost of locomotion (Telfer and Kelsall 1979,

26 1984) into areas with less snow cover, which results in an elevational migration from
27 highland to lowland ranges often observed in ungulates at northern latitudes (Myrsterud et al.
28 2011, Berg et al. 2019). Spatiotemporal variation in forage quality – as proposed by the
29 forage maturation hypothesis (FMH) – is a common explanation (Fryxell and Sinclair 1988,
30 Albon and Langvatn 1992, Hebblewhite et al. 2008) for individuals migrating away from the
31 winter area during summer. Migration, according to the FMH, is a tactic to exploit spatial
32 variation in plant phenology (or forage maturation) as the animals move along a gradient of
33 early plant growth, which provides high-quality forage, during spring green-up. Although
34 there is little doubt that ungulates move along the spring green-up wave (Bischof et al. 2012,
35 Van Moorter et al. 2013, Merkle et al. 2016, Aikens et al. 2017), it fails to explain why some
36 individuals do not follow this migration pattern in partially migratory populations.

37 The Competitive Release Hypothesis (CRH) suggests that certain individuals opt for a
38 different movement tactic to avoid intraspecific competition on the sympatric range
39 (Chapman et al 2011b, Berg et al. 2019). Van Moorter et al. (2020) presented a model for the
40 CRH based on density-dependent habitat selection. In general, density-dependent habitat
41 selection is a major mechanism resulting in an ‘ideal free distribution’ (IFD; Fretwell and
42 Lucas 1969, 1972). Under an IFD, individuals will experience equal fitness across patches
43 under a Nash equilibrium, i.e. moving from one patch to another will not increase individual
44 fitness. Van Moorter et al. (2020) showed that when the distribution of the most suitable
45 habitat changes across seasons, density-dependent habitat selection results in partial
46 migration. Moreover, they found that the migration rate should increase as the suitability
47 (intrinsic growth and/or carrying capacity) of the highland range increases compared to the
48 lowland range (Van Moorter et al. 2020).

49 In this study, we tested the predicted ideal free migration rates from the model
50 presented by Van Moorter et al. (2020) for migratory ungulates. These migrations are often

51 elevational from lowland ranges in winter to highland ranges in summer (e.g. Mysterud et al.
52 2011). We will use seasons (summer and winter) exclusively to indicate periods of the year
53 and adopt the geographic reference ('highland' and 'lowland' respectively) for the seasonal
54 ranges for simplicity and to avoid confusion. The model predicts that the proportion of
55 migrants increases as the relative habitat suitability of the highland range increases compared
56 to the lowland range (in agreement with the CRH; e.g. Chapman et al. 2011b, Berg et al.
57 2019). We define habitat suitability as habitat-dependent individual fitness, which is density-
58 dependent and corresponds to the intrinsic growth or carrying capacity in a habitat patch at
59 respectively low and high population densities (reviewed in Johnson 2007). In practice,
60 however, these two components are often highly correlated across habitats (Griffen and
61 Drake 2008 and references therein).

62 We tested the main prediction from the model using migration data from a large set of
63 GPS-tracked moose (*Alces alces*; $n = 545$) throughout Norway, Sweden, and Finland and
64 moose harvest statistics as an index of relative local population abundance, estimated at the
65 moose management area level. The tracking data, combined with remotely-sensed
66 environmental characteristics, allowed us to assess the suitability of highland and lowland
67 ranges for moose using resource selection functions (Manly et al. 2002, Boyce et al. 2002,
68 Boyce et al. 2016), which we validated using the harvest data. Vegetation and snow cover are
69 the most important factors determining seasonal ranges for moose (Ball, Nordengren and
70 Wallin, 2001, Allen et al. 2016). We expected a strong effect of snow (or a lack of) driving
71 seasonal habitat suitability: only the lowland range would be suitable habitat during winter,
72 whereas both the low- and highland range would be suitable during summer (baseline
73 Prediction 0). We then related local differences in the relative suitability of the highland
74 compared to the lowland range during summer to the proportion of migratory moose.
75 Following the CRH and more specifically our model predictions, we predicted that the

76 proportion of migratory moose would increase with both higher suitability of habitat
77 (Prediction 1) and overall greater availability of suitable habitat (Prediction 2) on the
78 highland range compared to the lowland range.

79

80 ***Material and Methods***

81 Model for partial migration

82 To investigate the relationship between seasonal variation in habitat suitability and the
83 proportion of migrants, we used the model presented by Van Moorter et al. (2020). This
84 model builds upon the two-patch population model from Fryxell and Holt (2013) and Holt
85 and Fryxell (2011), in which the dynamics of population size in each patch follow a Ricker
86 model (Ricker 1954, which is a discrete-time model); with scaled densities (*i.e.* $N = N'/K'$,
87 where N' is the unscaled population size in spring just prior to the calving period (May-June),
88 and K' is the carrying capacity during summer or population size at which each individual, on
89 average, replaces itself). Following Fryxell and Holt (2013), we assumed recruitment (r)
90 during summer to be density-dependent and mortality during winter (μ) to be density-
91 independent as supported by many empirical studies of ungulates (Saether 1997, Gaillard et
92 al. 2000). The two patches or seasonal ranges are linked through the movement of migratory
93 animals, which move with migration probability, m , and a demographic cost, c , after the
94 winter season from range L to H and back after summer. The range L is defined as the range
95 with the lowest winter mortality (*i.e.* $\mu_L \leq \mu_H$). See Van Moorter et al. (2020) for more details
96 (and the Supplementary Material for a summary).

97 The ideal-free migration strategy (Mariani et al., 2016; Van Moorter et al. 2020),
98 where the ideal-free migration probability (\hat{m}) equalizes the summer fitness for lowland
99 residents and migrants, is calculated as follows:

$$100 \quad \hat{m} = \frac{1 - \frac{r_L}{r_H} [1 - N] - \frac{c}{r_H}}{[\frac{r_L}{r_H} + \frac{K'_L}{K'_H}]N} \quad \text{Eq. 1}$$

101 This ideal-free migration probability is determined by three main components: migration cost
 102 (scaled by the recruitment on the highland range), relative habitat suitability of highland
 103 versus lowland range both in terms of the relative intrinsic growth ($\frac{r_H}{r_L}$) and the relative
 104 carrying capacity ($\frac{K'_H}{K'_L}$), and population density or saturation with respect to the lowland range
 105 (N). Consequently, increasing the cost of migration leads to a reduction in the migration
 106 probability. However, as the cost of migration seems negligible for moose (Rolandsen et al.
 107 2017), we will assume zero migration cost. Whereas, increasing the benefits of migration
 108 from access to suitable habitat (increasing $\frac{r_H}{r_L}$ or $\frac{K'_H}{K'_L}$) increases migration probability (Figure
 109 1). If the intrinsic growth rate differs between migrants and residents, then the migration rate
 110 will be density-dependent (Figure 1). If the intrinsic growth rate of migrants is higher than
 111 those of residents (e.g. due to higher quality forage following the forage maturation
 112 hypothesis; Fryxell and Sinclair 1988, Hebblewhite et al. 2008), all individuals of the
 113 population would be migrating at low densities. As the highland range becomes more
 114 crowded due to increasing population density, some individuals will shift to a resident
 115 strategy. The opposite will occur if the migrants experience a lower intrinsic growth on the
 116 highland ranges (red line in Figure 1), and then individuals will only start migrating once the
 117 population density on the lowland range is sufficiently high.

118

119 The study area

120 The study area covers nearly the whole of Fennoscandia: Norway, Sweden, and Finland (Fig.
 121 2). Climate varies greatly with latitude, elevation, and distance from the ocean. We find
 122 subarctic conditions in the north and more temperate conditions in the south. High elevation

123 areas are characterized by alpine conditions. Coastal areas are characterized by higher
124 precipitation than inland areas. The most common land cover class is coniferous forest,
125 especially in the north. Moors and heathland, and broad-leaved and mixed forests are also
126 commonly found in Fennoscandia. Even though most of the forested land is managed for
127 silviculture, human inhabited areas occupy only a small fraction of the land cover. Most areas
128 experience a snow free period of about 6 months, with longer snow cover and greater snow
129 depths in the north and at higher elevations.

130

131 Data

132 We analyzed both GPS-tracking data from adult moose (see Bunnefeld et al. 2010 for further
133 details) and harvest data from Finland, Norway, and Sweden, during 2005-2011. The GPS-
134 tracking data were collated in the Wireless Remote Animal Monitoring (Dettki et al. 2013)
135 database system for data validation and management. We screened these data using the
136 approach developed by Bjørneraas et al. (2010) and subsampled data to obtain one location
137 randomly throughout each day. We considered data from mid-January through mid-March as
138 winter data, and from July through the end of August as summer data, as moose during these
139 periods have usually finished their fall and spring migration, respectively (Rolandsen et al.
140 2010, Bunnefeld et al. 2010, Singh et al. 2012). All moose ($n = 545$) included in our analysis
141 had a complete summer and winter season. We used harvest statistics as a proxy for
142 abundance of moose at the moose management area level (Ueno et al 2014). Annual harvest
143 statistics (number of moose shot per year) were collected at municipality level in Norway
144 ($n=356$), hunting district level in Sweden ($n=308$), and county level in Finland ($n=60$).

145 We focused on two large-scale environmental variables: vegetation type or land cover
146 and snow cover. We used CORINE land cover data from 2006 (Büttner et al. 2004), which
147 has European coverage, including Finland, Norway, and Sweden. Snow cover is an important

148 environmental variable that restricts access to forage for moose during winter (except
149 conifers). Unfortunately, no common snow depth model was available for Fennoscandia, so
150 instead we used the duration of snow cover as a proxy. We followed a similar approach to
151 Dietz et al. (2015) to estimate the length of snow cover from MODIS remote-sensing
152 composites from an 8-day period (MOD10A2) with an approximately 500 m spatial
153 resolution. We linearly interpolated a maximum gap of 3 missing scenes (i.e. a gap of a
154 month). Using these data, we derived the average proportion of the year with snow cover
155 during the study period (2000-2011). The MODIS sensors require light to detect snow and
156 during the darkest period in winter no sun light is available above the arctic circle; we will
157 therefore likely underestimate the length of the snow season as we move north. However, as
158 few animals moved distances long enough to exploit latitudinal variation, we did not expect
159 this to affect our results.

160

161 Analysis

162 To quantify seasonal ranges (labeled 'lowland' and 'highland' to avoid confusion, see above),
163 we used seasonal resource selection probability functions (RSPF) instead of relying on the
164 actual geographic ranges occupied by moose in each season. We did this to avoid risking
165 circularity in our argument, as we relied on moose geographic space use to determine their
166 migratory status. The seasonal RSPF were based on a used and available points comparison
167 in environmental space, which comprises multiple dimensions representing different biotic
168 and abiotic environmental variables (Aarts et al. 2008). We assumed that moose in
169 Fennoscandia are generally well-adapted and select the most suitable locations for that season
170 from those available during winter and summer (i.e. those locations that give the highest
171 individual seasonal fitness with low winter mortality and high summer recruitment
172 respectively).

173 Testing the predictions from Equation 1 requires the quantification of habitat
174 suitability of the highland and lowland range during the summer season both in terms of
175 intrinsic growth rate and carrying capacity. These correspond to individual fitness at
176 respectively low and high population density. Griffen and Drake (2008) argue that the
177 intrinsic growth rate and carrying capacity refer to a population's response to the quality and
178 quantity of available habitat respectively. Indeed, in their experimental study on *Daphnia*
179 *magna*, Griffen and Drake (2008) found that both the quality and size of the habitat affect the
180 carrying capacity, whereas habitat quality mainly affects the intrinsic growth rate. We used
181 the selection probability of a habitat as an index of its quality or suitability (Manly 2002,
182 Boyce et al. 2002, Boyce et al. 2016), and tested this assumption using harvest statistics (see
183 below). The integrated (i.e. summed) selection probability over each range summarizes both
184 the quality and quantity of habitat available to moose, which we used as a proxy for the
185 carrying capacity. The maximum (i.e. 95 percentile) selection probability of each range
186 summarized the best available habitat, which we used as a proxy for the intrinsic growth rate.

187

188 *Habitat selection and Range prediction*

189 To identify the ranges that moose selected more during summer than during winter (i.e.
190 highland) or *vice versa* (i.e. lowland), we compared the selection probability in each
191 predicted pixel from the RSPF, which compares available and used habitat. We represented
192 the area available to each moose by the 99% Minimum Convex Polygon for the locations of
193 each individual with a 10 km buffer. For each GPS-location (i.e. used point) we sampled 10
194 points randomly from the available area to represent the habitat types available to each
195 moose. We estimated the maximum likelihood estimators of the RSPF with a logit-link
196 function from these use-available data using the combination of partial likelihood and data
197 cloning implemented in the ResourceSelection library (Lele and Keim, 2006; Lele, 2009) for

198 R (R Development Core Team, 2011). We analyzed the used locations from both seasons
 199 separately to obtain a population-level summer and winter RSPF, and the locations were
 200 matched with random points by individual to account for differential availability of habitat
 201 among individuals. We evaluated these RSPF using cross validation (similar to Boyce et al.
 202 2002) based on a leave-one-out approach. For this cross validation, we re-fitted the model to
 203 all but one individual and then predicted habitat use for this individual. We divided these
 204 predicted values into 10 equal sized bins, and then computed the spearman rank correlation
 205 between the bin number and the proportion of used versus available locations within each bin
 206 (see Boyce et al. 2002 for more details).

207 Following Lele et al. (2013): “The resource selection probability function, RSPF, [...] is defined as the probability that a resource unit of type x is selected (or, becomes part of the use set) when encountered.” As our analysis is focusing on regions where moose are
 208 generally present and moose have large ranging capabilities, we can assume all resource units
 209 to be available to moose. We therefore interpret our seasonal predictions as the higher
 210 probability of moose occurrence in a given location during winter or summer. We rescaled
 211 these probabilities to conserve the probability of moose occurrence over Fennoscandia across
 212 seasons.

213 We defined ‘lowland’ ranges as those pixels (x) that are more or equally selected (s)
 214 during winter than during summer (i.e. $s_s(x) \leq s_w(x)$), and the ‘highland’ ranges as those
 215 that are more selected during summer than during winter (i.e. $s_s(x) > s_w(x)$). We
 216 approximated the summer carrying capacity of these ranges using the integrated selection
 217 probability during the summer. Thus, we estimated the ratio of the summer and winter
 218 carrying (K) capacity as:

$$219 \frac{K_H}{K_L} = \frac{\sum_{x \in A} [s_s(x) > s_w(x)] s_s(x)}{\sum_{x \in A} [s_s(x) \leq s_w(x)] s_s(x)}$$

222 where numerator and denominator are the sum of selection probabilities during summer (s_s)
 223 for all pixels of the area available to a ‘population’ (A ; which is defined below) that are part
 224 of respectively its highland range ($s_s(x) > s_w(x)$) and its lowland range ($s_s(x) \leq s_w(x)$). The
 225 summer intrinsic growth rate (r) of these ranges was approximated using the 95 percentile of
 226 the selection probability during the summer:

$$227 \quad \frac{r_H}{r_L} = \frac{\text{perc}(s_s(x)) \text{ for } s_s(x) > s_w(x)}{\text{perc}(s_s(x)) \text{ for } s_s(x) \leq s_w(x)}$$

228

229 *Harvest statistics and Population distribution*

230 First, to test the appropriateness of the RSPF as a proxy for habitat suitability, we
 231 investigated the relationship between the average annual harvest (2005-2011) and the total
 232 RSPF for each reporting unit with linear regression. We log-transformed both harvest and
 233 summer and winter RSPF to reduce skew. Moose harvest in Fennoscandia takes place during
 234 fall, with most individuals harvested during September - October. As the main migration of
 235 moose to their winter range generally occurs later in the year (i.e. November-December; see
 236 Bunnefeld et al. 2010, Allen et al. 2016), most individuals will be harvested from their
 237 summer range. We therefore expected the summer RSPF to be a better predictor of harvest
 238 than the winter RSPF.

239 Second, to estimate the distribution (N') of moose within Fennoscandia, we
 240 distributed the mean number of harvested moose within each management unit in proportion
 241 to the summer RSPF for each pixel. The population density (N) is computed as the ratio of
 242 the abundance (N') and carrying capacity (K'), see the model description above. This density
 243 was computed for each ‘population’ cluster, see below.

244

245 *Proportion of migrants*

246 To quantify the proportion of migrants, we need first to identify the migratory status for each
247 individual, and second to group individuals into ‘population’ clusters. Following Cagnacci et
248 al. (2016), we assessed the migratory status of each individual using two methods: first, by
249 the overlap between seasonal home ranges, and second, by the movement distance between
250 them. Overlap in seasonal home ranges was determined using the volume of intersection for
251 the kernel utilization distributions (Fieberg and Kochanny 2005) from winter and summer
252 locations (respectively: January 15th till March 15th and July 1st till August 31st). We
253 considered migrating animals to be those that had less than 1% overlap between their
254 seasonal home ranges. We used `adehabitatHR` library (Calenge 2006) for R (R Development
255 Core Team, 2011) to calculate seasonal home range overlap. We based the second migration
256 criterion on the straight-line distance between the median summer and the median winter
257 locations. We considered animals with migration distances >10 km as migratory animals,
258 whereas animals with shorter distances were considered resident (see Hjeljord 2001).

259 Second, although moose throughout Scandinavia experience different local
260 conditions, there are no distinct boundaries between different ‘populations’. To analyze the
261 proportion of migrants as a function of local environmental conditions, we therefore
262 performed a clustering analysis on the median winter location of each moose. We used
263 hierarchical clustering (‘`hclust`’) with the gap-statistic to identify the optimal number of
264 clusters (Tibshirani et al., 2001) in R with the `factoextra` library. We do not assume these
265 clusters to correspond to separate sub-populations, they merely represent clusters of moose
266 that experience similar local conditions.

267 We then computed the proportion of migrants for each cluster using either the overlap
268 or distance criterion between seasonal home ranges. We considered the area within 300 km of
269 a cluster to be available to the moose within that cluster, with this distance combining both
270 the long-distance migrations observed in moose (up to about 200 km, see Bunnefeld et al.

271 2010) and the spatial variation in location of moose within each cluster. For this area, we
 272 computed the ratio of the highland and lowland range intrinsic growth and carrying capacity
 273 as explained above. The density of moose for each cluster was computed as the sum of the
 274 moose distribution within the area (derived from the annual harvest, see above) divided by
 275 the carrying capacity.

276

277 *Hypothesis testing*

278 Finally, to test our hypotheses, we fitted four statistical models to the data. First, a basic
 279 logistic regression used the range ratios as predictors of the proportion of migrants. However,
 280 as the correlation between the ratio of the intrinsic growth rate and carrying capacity was
 281 very high ($r=0.97$), we did not include them in the same model:

$$282 \quad m = \beta_0 + \beta_1 \frac{r_H}{r_L} + \varepsilon \quad \text{Eq. 2a}$$

$$283 \quad m = \beta_0 + \beta_1 \frac{K_H}{K_L} + \varepsilon \quad \text{Eq. 2b}$$

284 Second, using nonlinear least-squares estimates ('nls' from the nlme-library [Pineiro et al.
 285 2012] in R [R Development Core Team, 2011]), we parameterized Equation 1, assuming no
 286 cost to migration:

$$287 \quad m = \frac{1 - \beta_1 \frac{r_L}{r_H} (1 - \beta_3 N)}{\left(\beta_1 \frac{r_L}{r_H} + \beta_2 \frac{K_L}{K_H} \right) (\beta_3 N)} \quad \text{Eq. 3}$$

288 here β_1 estimates the effect of the ratio of intrinsic growth rates between lowland and
 289 highland range, β_2 the effect of the ratio of carrying capacity on both ranges, and β_3 the
 290 effect of the density of moose in Fennoscandia on the lowland ranges during summer. We
 291 expected $\beta_1 = \beta_2 = 1$ according to Equation 1, whereas β_3 represents the proportional
 292 constant between our proxy for density and the actual population density.

293 Third, given the coarseness of our density proxy, we also fitted a simplified model
 294 where we assumed a constant density across Fennoscandia:

$$295 \quad m = \frac{1 - \beta_1 \frac{r_L}{r_H} (1 - \beta_3)}{\left(\beta_1 \frac{r_L}{r_H} + \beta_2 \frac{K_L}{K_H} \right) \beta_3} \quad \text{Eq. 4}$$

296 Finally, we further simplified the model by assuming both $N=1$ and $\frac{r_L}{r_H} = 1$ to
 297 focus exclusively on the ratio between the highland and lowland range of availability of
 298 suitable habitat:

$$299 \quad m = \frac{1}{\beta_1 + \beta_2 \frac{K_L}{K_H}} \quad \text{Eq. 5}$$

300

301 **Results**

302 *Habitat selection and Range prediction*

303 Duration of the snow cover season and the land cover classes affected moose habitat selection
 304 both in summer and winter. As expected (Prediction 0), the duration of the snow-free period
 305 had a stronger effect on space use during winter (slope \pm SE: 12.23 ± 0.43) than during
 306 summer (slope \pm SE: -2.55 ± 0.14). In addition, considerable differences existed in the
 307 selection of the land cover classes. Moose avoided most other land cover classes compared to
 308 coniferous forests, and selected only a few land cover classes over it (Table 1). In summer,
 309 moose selected for broad-leaved forests over coniferous forests, whereas broad-leaved forests
 310 were avoided during winter. A similar effect, albeit much weaker, was found for mixed
 311 forests, which moose selected during summer and avoided during winter compared to
 312 coniferous forests. In winter, woodland-shrub tended to be strongly selected over coniferous
 313 forests, a selection that was much weaker in effect during summer. The habitat selection
 314 models showed high cross validation performance based on the leave-one-out approach (see
 315 Supplementary Figure S2), both during summer (correlation: median, mean \pm SE: 0.53, 0.45

316 ± 0.02) and during winter (correlation: median, mean \pm SE: 0.64, 0.56 ± 0.01). We used these
317 habitat selection models to predict the preferred ‘highland’ and ‘lowland’ ranges (see Figure
318 2), which as expected occur on average at different elevations (average elevation on the
319 lowland and highland range was respectively 174 m and 401 m, p -value < 0.001 ; see
320 Supplementary Figure S6).

321

322 *Harvest and Habitat suitability*

323 We found a positive relationship between harvest and the RSPF (see Figure 3), which
324 supported our interpretation of the RSPF as a metric for habitat suitability. Somewhat
325 surprisingly, we found a lower AIC for the RSPF from winter than from summer (Δ
326 $AIC = 6.6$). The relationship between the RSPF and harvest was close to unity on a
327 logarithmic scale (winter: $\beta \pm SE$: 1.05 ± 0.04 ; summer: $\beta \pm SE$: 0.95 ± 0.04), with half the
328 variance in harvest explained by the RSPF (for either the winter or summer RSPF). A notable
329 deviation from the predicted relationship between harvest and RSPF occurred in Western
330 Norway (see Supplementary Figure S3), where, despite the presence of suitable moose
331 habitat, harvest is close or equal to zero. This is probably due to a combination of historical
332 overharvesting, movement barriers, and competition with local red deer populations. As these
333 areas are not occupied by a sizable moose population, they are not part of our study area on
334 moose migration. Removing these areas would result in a stronger relationship between
335 harvest and RSPF (see Supplementary Figure S4), with the harvest better predicted by the
336 summer than winter RSPF ($\Delta AIC = 236.5$), as expected.

337

338 *Migration probability*

339 When clustering individuals into groups experiencing a similar environment, the gap-statistic
340 kept increasing up to 37 clusters, with extremely small cluster sizes. We therefore opted to

341 apply the ‘elbow’ criterion by selecting the number of clusters where the increase in gap-
342 statistic decelerated. We clustered the moose into 19 clusters, where moose within a cluster
343 were on average 35 km apart (and up to 165 km).

344 Our two migration criteria, based on the overlap versus the distance between winter
345 and summer home ranges (see above for a detailed description), gave very similar results for
346 each individual: there was 85% agreement between methods. The proportion of migrants in
347 each cluster was highly correlated ($r > 0.95$, see Supplementary Figure S8), giving
348 qualitatively identical results. We therefore only report the results from the range overlap
349 criterion.

350 We found a strong positive correlation between the ratio of the intrinsic population
351 growth rate and the ratio of the carrying capacity between highland and lowland ranges
352 ($r = 0.96$, $p < 0.001$); therefore, we did not use both variables simultaneously in the logistic
353 regression. We found a significant positive effect of both ratios on the proportion of migrants
354 (Table 2; ratio of intrinsic growth: $\beta_1 \pm \text{SE}$: 0.85 ± 0.14 , $p\text{-value} < 0.001$; ratio of carrying
355 capacity: $\beta_1 \pm \text{SE}$: 0.36 ± 0.07 , $p\text{-value} < 0.001$). The AIC was lower for the model explaining
356 the proportion of migrants based on the proxy for intrinsic growth than for the proxy for
357 carrying capacity ($\Delta\text{AIC} = 6$).

358 Estimating the parameters in Equation 1 (assuming cost $c=0$), we found that the best
359 model was the one in which we made the additional assumptions of constant intrinsic growth
360 rate ratios and the population density to be one across clusters (Eq. 5, Table 2). The fully
361 parameterized model (Eq. 3) suffered from convergence issues, while after constraining the
362 parameter range to realistic values, some estimates corresponded to those boundaries (see
363 Table 2). Thus, the estimates for Eq. 3 are unreliable. Assuming constant density across
364 clusters (Eq. 4) allowed the model to converge with a density of one (i.e. high population
365 density).

366 In the best model (i.e. Eq. 5), the intrinsic growth rates on the highland and lowland
367 ranges were not significantly different (Table 2; $\beta_1 (\pm SE)$: $1.37 (\pm 0.33) \approx 1$). As expected
368 from Equation 1, the effect of the ratio of the carrying capacity between the lowland and
369 highland ranges was not significantly different from one (Table 2 and Figure 3; $\beta_2 \pm SE$: 0.98
370 ± 0.68).

371

372 ***Discussion***

373 Multiple explanations have been proposed for the evolution of differential migratory
374 tendencies among individuals within a population (reviewed in Chapman et al. 2011b and
375 Berg et al. 2019), but there is still lack of a quantitative framework to predict the level of
376 partial migration. Using a simple density-dependent habitat selection model from Van
377 Moorter et al. (2020), we propose (Eq. 1 and Figure 1) that the ideal-free migration
378 probability is determined by three main components: migration cost (scaled by the
379 recruitment on the highland range), relative habitat suitability of the seasonal range (i.e. $\frac{r_H}{r_L}$
380 and $\frac{K'_H}{K'_L}$), and population density or saturation (N). Given its basis in ideal-free migration, the
381 proposed model presents a formal representation of the Competitive Release Hypothesis, as
382 individuals opt for a different strategy to escape from competition on the sympatric range
383 (e.g. Berg et al. 2019). In our empirical test of this model, we found that moose ranges in
384 Fennoscandia do indeed show an expansion-contraction dynamic during summer and winter
385 (Prediction 0), and that the proportion of migrants increased with the amount of suitable
386 habitat becoming available during summer (Prediction 2). However, the proportion of
387 migrants did not increase with their access to higher quality habitat *per se* (Prediction 1).

388 *Seasonal Range Dynamics*

389 We found seasonal changes in moose's selection of land cover classes, such as an increased
390 selection for broad-leaved forest, natural grassland, moors and heathland during the summer.
391 However, the main driver for the seasonal range expansion-contraction dynamic was the
392 difference in duration of the snow-free period due to elevational gradients. The link between
393 the almost universal seasonal range contraction of cervids in fall and snow has been known
394 for decades (e.g. Brazda 1953, LeResche 1974). Surprisingly, an underlying explanation of
395 range expansion in spring including the mechanism of density-dependent habitat selection –
396 where partial migration is a competition avoidance tactic (Kaitala et al. 1993, Taylor and
397 Norris 2007, Holt and Fryxell 2011) – has rarely been focused on. In our case, this is
398 addressed in the ratio of the suitability of the highland range relative to the lowland range.

399 In our tests of the model predictions, we found support for the expected increase in
400 migration probability as the relative high-density suitability of the highland range increased
401 compared to the lowland range ($\frac{K'_H}{K'_L}$; in support of Prediction 2). Moreover, the observed
402 increase was not significantly different from the model prediction (observed slope \approx predicted
403 slope = 1). However, our best model did not lend support to the relative low-density
404 suitability of highland versus lowland range ($\frac{r_H}{r_L}$) as a driver for migration (not supporting
405 Prediction 1). This absence of an effect of the intrinsic growth rate could be due to (1) our
406 proxy not adequately capturing the spatial heterogeneity in intrinsic growth, or (2) the similar
407 response of intrinsic growth and carrying capacity to spatial heterogeneity, which resulted in
408 highly correlated metrics. Although different from a theoretical perspective, intrinsic growth
409 and carrying capacity respond relatively similarly to changes in habitat quality and quantity
410 (Griffen and Drake 2008). Alternatively, (3) the role of variation in intrinsic growth rate may
411 be dependent upon the level of population saturation on the lowland range during summer, as
412 predicted by the model (Figure 1). Moose densities in Fennoscandia are generally high

413 (Lavsund et al. 2003, Jensen et al. 2020), which could lead to the intrinsic growth rate not
414 being that important for migration in Fennoscandia.

415 *Population Density and Habitat Suitability*

416 Previous studies have found conflicting roles of increasing population density on the
417 proportion of migrants in partially migratory populations. For instance, Geremia et al. (2011)
418 found an increase in the proportion of bison in Yellowstone that migrated outside the park,
419 whereas Mysterud et al. (2011) reported a decrease in the number of migratory red deer
420 (*Cervus elaphus*), with an increase in density. Part of this discrepancy could be due to using a
421 spatial contrast in density rather than temporal variation in density in the latter case, and the
422 link to competition therefore being uncertain. However, migrants' access to high quality
423 forage (following the Forage Maturation Hypothesis) would lead to migration in spring even
424 at low population density (Figure 1; Mysterud et al. 2012), which could explain a reverse
425 density-dependence with a reduced proportion of migrants until high elevation ranges fill up
426 at high density. In the model (Eq.1), whether migration rates increase or decrease with
427 population density depends upon the intrinsic quality of the lowland compared to the
428 highland range (Fig. 1). At low densities, all individuals would migrate towards an
429 intrinsically better highland range or remain resident in an intrinsically better lowland range.
430 Whereas, as densities increase, an increasing proportion of individuals would adopt the
431 opposite tactic to get released from competition. Thus, contrasting effects of density on the
432 migration rate are predicted by this model.

433 In our study, we did not find support for an effect of spatial variation in local moose
434 densities on the migration rate (as expected from Equation 1). However, the proportion of
435 migrants is most sensitive to changes in population density at low densities and when the
436 intrinsic growth rate differs between the highland and lowland ranges (Figure 1). We found

437 no support for an effect of differences in intrinsic population growth on proportion of
438 migratory moose, and moose densities in Fennoscandia are generally relatively high
439 (Lavsund et al. 2003, Jensen et al. 2020).

440 *Limitations and Further Developments*

441 As with all models, we made several simplifying assumptions (cf. Holt and Fryxell 2011;
442 Fryxell and Holt 2013; Van Moorter et al. 2020). For instance, our model does not include
443 ecological interactions other than immediate density-mediated competition during the
444 summer season (see also Fryxell and Holt, 2013). When allowing for density-dependent
445 competition during both seasons, as in the original Holt and Fryxell (2011) model, no simple
446 solution exists for the proportion of migrants and more complex dynamics may occur (Steinar
447 Engen, pers. comm.).

448 To avoid confusion with seasons as periods of the year, we referred to the winter
449 range and exclusive summer range as respectively the “lowland” and “highland” range. This
450 was a convenient terminology because the seasonal range dynamics were largely driven by
451 differences in the duration of the snow-free period linked to an elevational gradient, as is
452 often observed with ungulates at northern latitudes (Myrsterud et al. 2011). Although, while
453 the “lowland” ranges on average were at lower elevation than the “highland” ranges (see also
454 Supplementary Figure S6), it should be noted that deviations from this elevational pattern
455 occur as the choice of seasonal range is not only driven by snow cover (Ball et al. 2001). We
456 found differences in land cover classes between the seasonal ranges. Previous studies have
457 also documented marked differences in forest composition between summer and winter
458 ranges linked to seasonal shifts in the diet (Histøl and Hjeljord 1993, Månsson et al. 2007,
459 Wam and Hjeljord, 2010). This may also explain why moose may choose an opposite
460 strategy where they migrate towards higher elevation winter areas (Andersen 1991). Thus, the

461 terms lowland and highland range in this paper refer to areas that are respectively more or
462 less preferred during winter than during summer, rather than strictly the elevation of those
463 areas.

464 For browsing herbivores, like moose, one expects a lagged decrease in forage
465 availability on the lowland range after the presence of many animals during winter (Persson,
466 Danell, and Bergström 2005; van Beest et al. 2010). Illius and O'Connor (2000) explored the
467 ecological effects of range expansion and contraction due to seasonal rainfall in semi-arid
468 grazing systems. They found that increased range contraction leads to increased pressure on
469 the contracted area and decreased pressure on the expansion area. Such lagged effects of
470 density were not implemented in our model and may reduce carrying capacity and intrinsic
471 growth rate on the lowland range during summer, and could lead to a higher migration
472 probability.

473 Our model did not include individual heterogeneity such as age, which are known to
474 influence migration probability in moose, or unequal competition between migrants and
475 residents on the lowland range (Histøl and Hjeljord 1993, Singh et al. 2012). Several
476 explanations for partial migration are based on individual heterogeneity (see Chapman et al.
477 2011b). The transition equations (e.g. Supplementary Material Eq. S2) could easily be
478 extended to include individual heterogeneity, which is an interesting avenue for future
479 research. Although, some variation in body size between resident and migrant individuals has
480 been observed in moose (Rolandsen et al. 2017), this was considered a consequence rather
481 than the cause of their different migratory behavior.

482 The assumption of habitat selection as a proxy for habitat suitability is not trivial (e.g.
483 Gaillard et al. 2010). Remotely-sensed data such as length of snow cover are relatively crude
484 proxies for ecological dynamics and so are land use – land cover maps. For instance, moose

485 are likely more affected by the conditions (wet vs. dry) and depth of the snow than by the
486 cover *per se* (Ball et al. 2001, Singh et al. 2012). In addition, the habitat selection process
487 itself may also be more complex than addressed in a simple RSPF, for instance, selection for
488 habitat features may depend upon population density (van Beest et al. 2013, Avgar et al.
489 2020). Fortunately, the strong positive relationship we found between the annual moose
490 harvest in areas with more selected habitat (Figure 2) is supportive of this assumption in our
491 study, indicating that the RSPF also represents the main dynamics affecting habitat
492 suitability.

493 Finally, we did not address the cost of migration and assumed it was zero in our
494 analysis. In our study system, the demographic cost of migration, if there is any, seems
495 extremely low (Rolandsen et al. 2017), as given by the extremely low natural mortality rate
496 of adult moose in Fennoscandia (Solberg et al. 2005). Also, the pure energetic cost associated
497 with migration is likely rather limited for moose, as their migrations seem to result more from
498 their movements becoming more directional, than by a strong increase in their movement rate
499 (Van Moorter et al. 2013).

500

501 *Conclusions*

502 Our results show that spatial variation in the proportion of migrants is partly explained by
503 spatial variation in seasonal changes in suitable habitat, which results in range expansion-
504 contraction. This can contribute to an improved management and conservation of partially
505 migratory species (Allen and Singh 2016), which is crucially needed as migratory behavior in
506 large ungulates is under pressure from human fragmentation and climate change (Berger
507 2004, Bolger et al. 2008). Migratory species can range over larger areas than administrative
508 units of management (Meisingset et al. 2018), which complicates their management

509 substantially (Thirgood et al. 2004, Skonhøft 2005). For example, Nilsen et al. (2009) showed
510 that migratory moose can cause a dissociation of costs (i.e. browsing damage and moose-
511 vehicle collisions) and benefits (i.e. harvest) associated with moose, leading to non-optimal
512 management. Therefore, areas with larger range expansion-contraction face more challenges
513 in reaching management and conservation goals for partially migratory species (Allen et al.
514 2016). Moreover, as changes in snow cover due to climate change are likely to affect the
515 benefits of migration at northern latitudes, our framework can be extended to predict
516 developments of partial migration under climate change.

517

518 **References**

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723 TABLES

724 Table 1

725 Parameter estimates for moose Resource Selection Functions in summer and winter. The
 726 reference land cover class was coniferous forest, which is the most common land cover class
 727 in Fennoscandia.

	Summer		Winter	
	$\beta \pm SE$	<i>p</i> -value	$\beta \pm SE$	<i>p</i> -value
Intercept	-1.98 ± 0.99	< 0.001	-7.87 ± 0.18	< 0.001
Length snow-free period	-2.55 ± 0.14	< 0.001	12.23 ± 0.43	< 0.001
Artificial structures	-2.81 ± 0.22	< 0.001	-3.68 ± 0.15	< 0.001
Homogeneous agriculture	-0.86 ± 0.04	< 0.001	-1.88 ± 0.07	< 0.001
Heterogeneous agriculture	-0.31 ± 0.03	< 0.001	-0.52 ± 0.05	< 0.001
Broad-leaved forest	1.23 ± 0.05	< 0.001	-1.83 ± 0.04	< 0.001
Mixed forest	0.41 ± 0.03	< 0.001	-0.10 ± 0.04	< 0.01
Natural grassland	-0.77 ± 0.12	< 0.001	-2.26 ± 0.26	< 0.001
Moors and heathland	-0.49 ± 0.03	< 0.001	-3.13 ± 0.05	< 0.001
Woodland shrub	0.35 ± 0.02	< 0.001	1.32 ± 0.04	< 0.001
Sparse vegetation	-2.00 ± 0.06	< 0.001	-2.75 ± 0.09	< 0.001
Wetlands	-0.07 ± 0.03	< 0.005	-1.16 ± 0.04	< 0.001
Water bodies	-2.63 ± 0.08	< 0.001	-4.76 ± 0.13	< 0.001
Unclassified	0.94 ± 0.08	< 0.001	-3.31 ± 0.50	< 0.001

728

729 Table 2

730 Parameter estimates from the analyses of the proportion of moose migrants (migrants had less
 731 than 1% overlap between the summer and winter home ranges). We performed five analyses:
 732 ordinary logistic regression using intrinsic growth rate ratio (Eq. 2a in the main text) or
 733 carrying capacity ratio (Eq. 2b), and non-linear least squares parameterization of Eq. 1 in the
 734 main text based on three sets of additional assumptions (corresponding to Eq. 3, 4, and 5 in
 735 the main text). See the main text for further details.

		AIC:	224
Eq. 2a	$\beta_0 + \beta_1 r_H / r_L$	$\beta_0 \pm \text{SE}$:	-1.60 ± 0.28
		$\beta_1 \pm \text{SE}$:	0.85 ± 0.14
		AIC:	230
Eq. 2b	$\beta_0 + \beta_1 K_H / K_L$	$\beta_0 \pm \text{SE}$:	-0.81 ± 0.18
		$\beta_1 \pm \text{SE}$:	0.36 ± 0.07
		AIC:	15
Eq. 3	$\frac{1 - \beta_1 r_L / r_H (1 - \beta_3 N)}{(\beta_1 r_L / r_H + \beta_2 K_L / K_H) (\beta_3 N)}$	$\beta_1 \pm \text{SE}$:	1.11 ± 1.28
		$\beta_2 \pm \text{SE}$:	$5.00^* \pm 5.51$
		$\beta_3 \pm \text{SE}$:	$20.00^* \pm 19.47$
		AIC:	6
Eq. 4	$\frac{1 - \beta_1 r_L / r_H (1 - \beta_3)}{(\beta_1 r_L / r_H + \beta_2 K_L / K_H) \beta_3}$	$\beta_1 \pm \text{SE}$:	3.88 ± 1.44
		$\beta_2 \pm \text{SE}$:	0.06 ± 2.37
		$\beta_3 \pm \text{SE}$:	$1.00^* \pm 0.28$
		AIC:	-0.4
Eq. 5	$\frac{1}{\beta_1 + \beta_2 K_L / K_H}$	$\beta_1 \pm \text{SE}$:	1.37 ± 0.33
		$\beta_2 \pm \text{SE}$:	0.98 ± 0.68

736 * Note: these estimates resulted in the upper boundary set for the parameter search, without
 737 these upper boundaries the model did not converge.

738

739 FIGURES

740 **Figure captions**

741 Figure 1

742 The proportion of moose migrants as a function of population density, growth rate, and
743 carrying capacity. Following Equation 1, the proportion of migrants depends on the migration
744 cost (here assumed zero), population density on the lowland range, the ratio of the lowland
745 and highland range growth rates (lines in red = 1.05, in black = 1, in light blue = 0.95, and in
746 dark blue = 0.9), and the ratio of the lowland and highland range carrying capacity (solid
747 lines = 2, dashed = 1, dot-dashed = 0.5, and dotted lines = 0.33). Note, for clarity we only
748 displayed the $r_L/r_H = 1.05$ (red line) for the low carrying capacity ratio ($K_L/K_H = 2$). See main
749 text for further discussion.

750

751 Figure 2

752 The seasonal ranges and the proportion of moose migrants (black) versus residents (grey) for
753 19 clusters in Fennoscandia. The lowland range is shown in blue and the highland range in
754 green. See Supplementary Figure S7 for an overview of the number, sex, and age of the
755 individuals in each cluster.

756

757 Figure 3

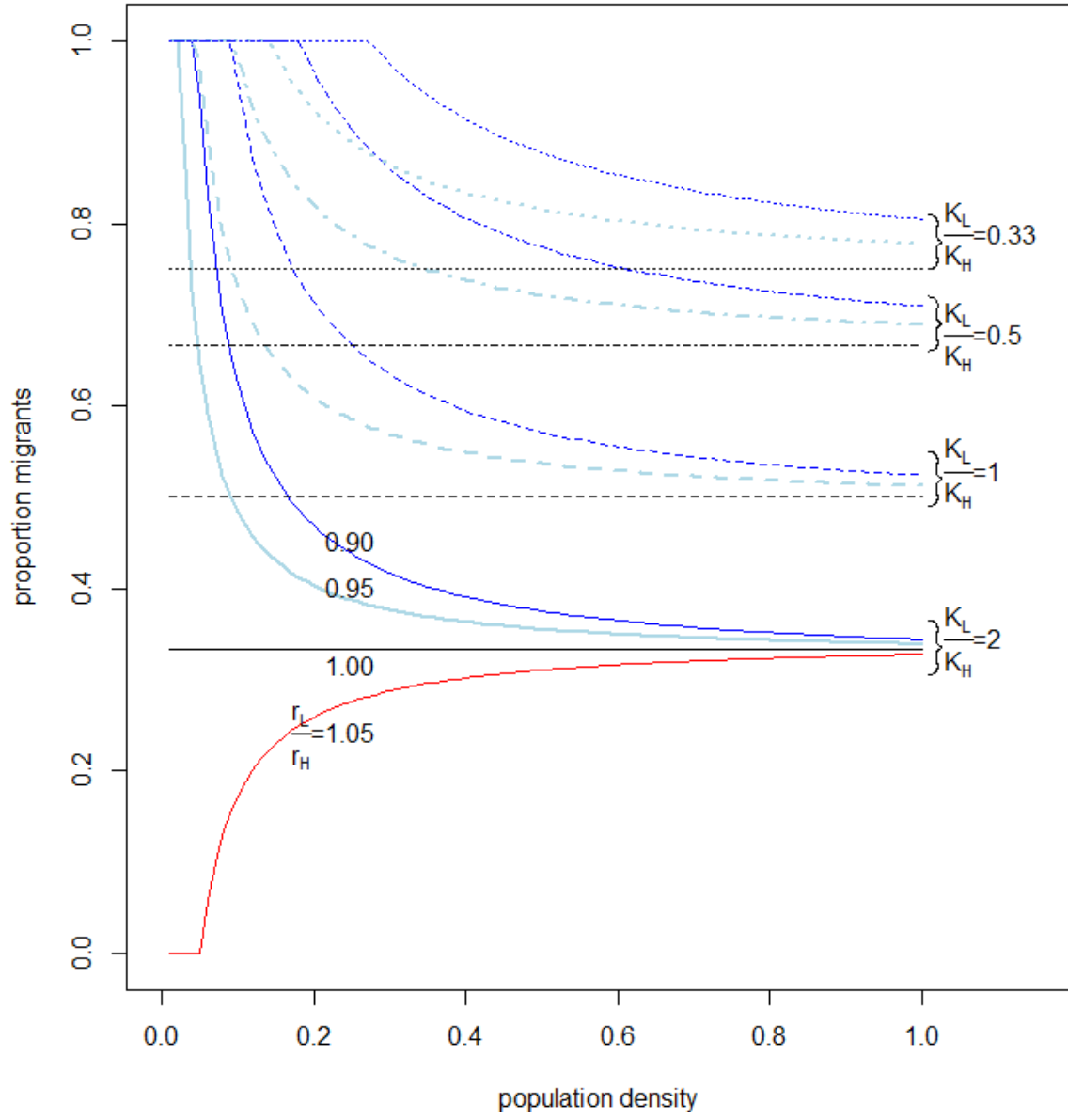
758 The relationship between moose harvest and habitat suitability in Fennoscandia. The
759 logarithm of the average annual harvest plotted against the total summer suitability for each
760 reporting unit in Fennoscandia with the fitted regression line.

761

762 Figure 4

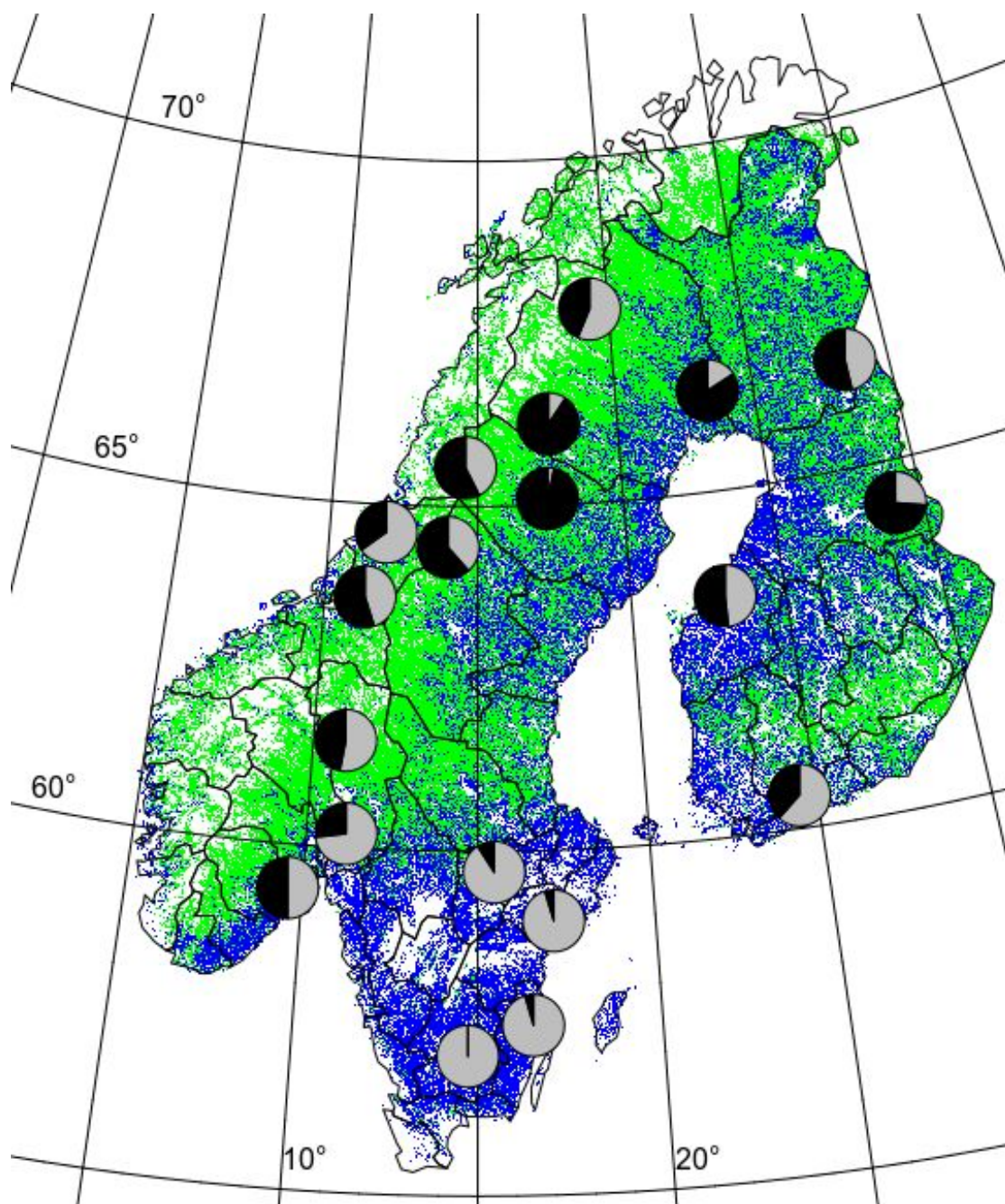
763 The proportion of migrants versus the size of the ratio in highland and lowland range carrying
764 capacity for each cluster. The proportion of migrants is calculated using the seasonal range
765 overlap criterion. The clusters are marked with the first letter of the country (i.e. Finland,
766 Norway, and Sweden) from which most moose originated. The full grey line represents the
767 expected proportion of migrants following Eq. 1, when both ranges would not differ in
768 intrinsic growth ($r_L/r_H = 1$), are fully saturated ($N_w=I$), and migration is cost-free ($c=0$).
769 The different fitted models from Eq. 2b, 3, 4, and 5 are represented respectively by a dashed,
770 dot-dashed, dotted, and full black line. See Table 1 for the model performance and parameter
771 estimates, and the main text for further explanations.

773 **Figure 1**



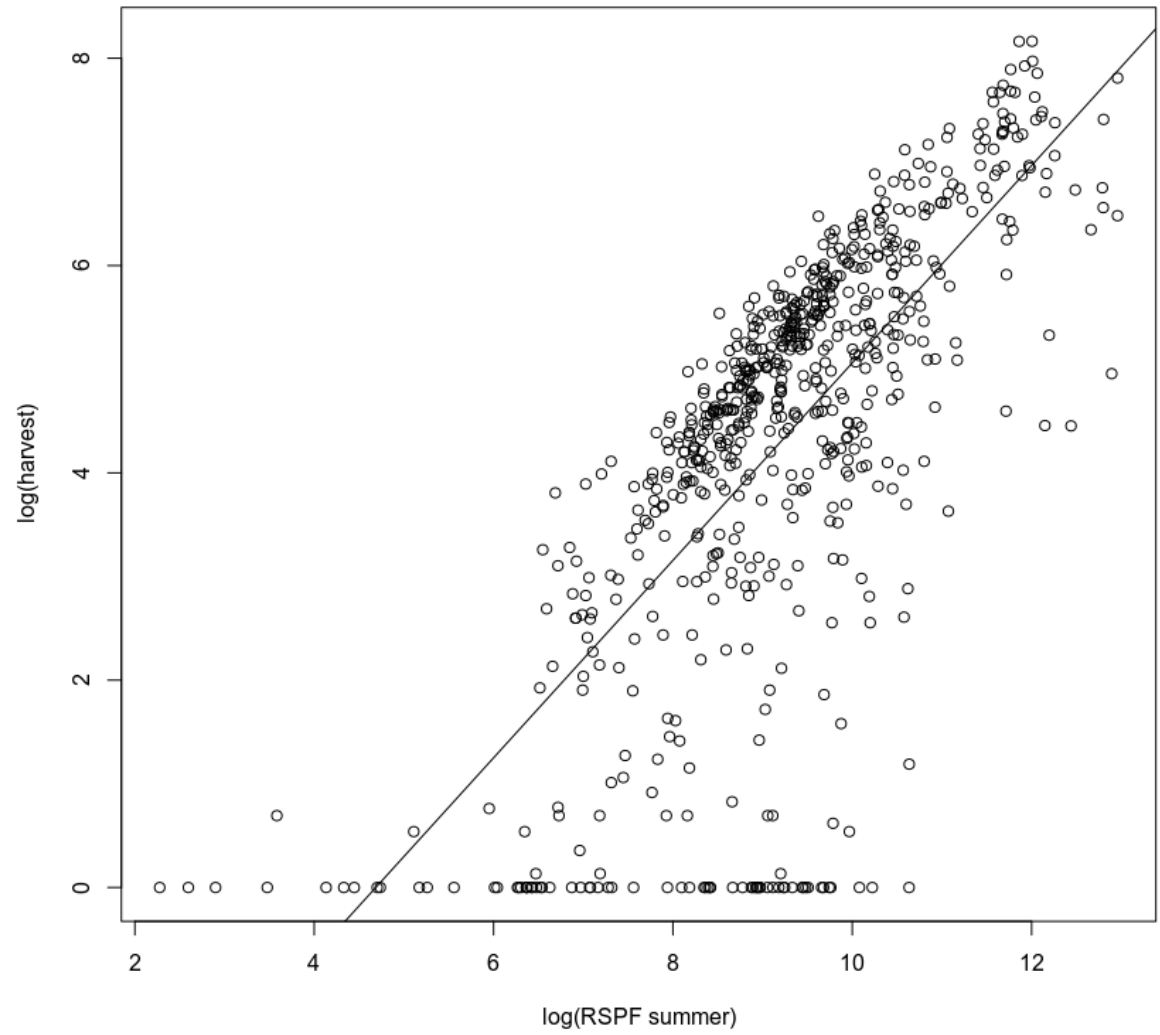
774

775 **Figure 2**



776

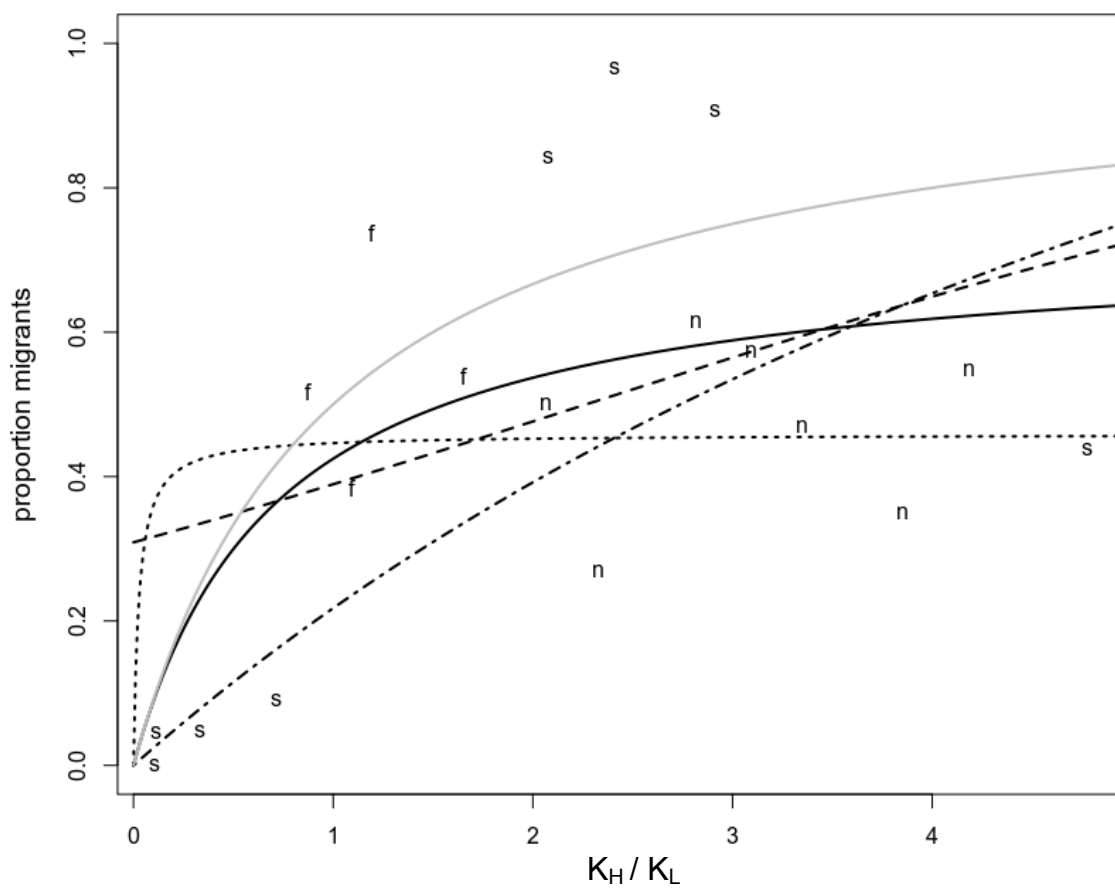
778 **Figure 3**



779



780 **Figure 4**



781

only

Supplementary Material

Model for partial migration

To investigate the relationship between seasonal variation in habitat suitability and the proportion of migrations, we used the model presented by Van Moorter et al. (2020). This model builds upon the population model from Fryxell and Holt (2013), we refer to Van Moorter et al. in which the dynamics of population size follow a Ricker model (Ricker 1954, which is a discrete-time model); with scaled densities (*i.e.* $N = N'/K'$, where N' is the unscaled population size in spring just prior to the reproduction season, and K' is the population size at which on average each individual replaces itself during summer). Following Fryxell and Holt (2013), we assumed recruitment (r) during summer to be density-dependent and mortality during winter (μ) to be density-independent as supported by many empirical studies of ungulates (Saether 1997, Gaillard et al. 2000).

Using the Ricker formula to represent episodes of summer reproduction, the multiplicative growth rate equals $\exp(r[1-N'/K']) = \exp(r[1-N])$, where e^r is the maximum per capita recruitment during summer. Assuming density-independent winter survival probability $e^{-\mu}$, the number of animals after one year at the end of winter is calculated as follows:

$$N(t+1) = N(t) \exp(r[1 - N(t)] - \mu) \quad \text{Eq. S1}$$

Following Fryxell and Holt (2013), we linked two seasonal ranges through the movement of migratory animals, which move with migration probability, m , and cost, c , after the winter season from range L to H and back after summer. We define range L as the range with the lowest winter mortality (*i.e.* $\mu_L \leq \mu_H$). Therefore, we assumed only migration during spring from L to H; we did not consider individuals moving in the opposite direction, the so called 'perverse' migrants (*sensu* Fryxell and Holt, 2013). Importantly, we focus our analysis on migration from the shared lowland range in winter towards a highland range used only in summer, *i.e.* the highland range can either not sustain residents year-round or highland residents are outcompeted by migrants who benefit from lower mortality on the lowland range (see for more details Van Moorter et al. 2020).

Following the population model in Equation S1, the number of animals at the end of winter in the lowland range, N , after 1 year is the sum of residents in L and migrants:

$$N(t+1) = (1 - m)N(t) \exp(r_L [1 - (1 - m)N(t)] - \mu_L) + mN(t) \exp(r_H [1 - \frac{K'_L}{K'_H} mN(t)] - \mu_L - c) \quad \text{Eq. S2}$$

where m is the migration probability, and c is the demographic cost of migration. For simplicity, we assumed that the migration cost was incurred after summer. Density-dependent recruitment (r) takes place during summer, which is for migrants in range H (*i.e.* r_H) and for lowland residents in range L (*i.e.* r_L), whereas the density-independent winter mortality (μ) occurs when migrants and lowland range residents share their common lowland range L.

As more animals migrate (*i.e.* increasing migration probability, m) from range L, the density of animals in L decreases and the summer fitness of residents in L increases, whereas the density of animals in H increases and the summer fitness of migrants in L decreases. We assumed an ideal-free migration strategy (Mariani et al., 2016; Van Moorter et al. 2020), where the ideal-free migration probability (\hat{m}) equalizes the summer fitness for lowland residents and migrants, calculated from Equation S2:

$$\exp(r_L [1 - (1 - m)N(t)]) = \exp(r_H [1 - \frac{K'_L}{K'_H} mN(t)] - c)$$

Since this equation is linear in m , the ideal-free migration probability (\hat{m}) is calculated as follows:

$$\hat{m} = \frac{1 - \frac{r_L}{r_H} [1 - N] - \frac{c}{r_H}}{[\frac{r_L}{r_H} + \frac{K'_L}{K'_H}]N} \quad \text{Eq. S3 = Eq. 1 in main text}$$

This ideal-free migration probability is determined by three main components: migration cost (scaled by the recruitment on the highland range), relative habitat suitability of highland versus lowland range both in terms of the relative intrinsic growth ($\frac{r_H}{r_L}$) and the relative carrying capacity ($\frac{K'_H}{K'_L}$), and population density or saturation with respect to the lowland range (N). Consequently, increasing the cost of migration leads to a reduction in the migration probability. However, as the cost of migration seems negligible for moose (Rolandsen et al. 2017), we will assume zero migration cost in this paper. Whereas, increasing the benefits of migration from access to suitable habitat (increasing $\frac{r_H}{r_L}$ or $\frac{K'_H}{K'_L}$) increases migration probability (Figure 1). If the intrinsic growth rate differs between migrants and residents, then the migration rate will be density-dependent (Figure 1). If the intrinsic growth rate of migrants is higher than those of residents (e.g. due to higher quality forage following the forage maturation hypothesis; Fryxell and Sinclair 1988, Hebblewhite et al. 2008), all individuals of the population would be migrating at low densities. As the highland range becomes more crowded due to increasing population density, some individuals will shift to a resident strategy. The opposite will occur if the migrants experience a lower intrinsic growth on the highland ranges (red line in Figure 1), then individuals will only start migrating once the population density on the lowland range is sufficiently high.

Supplementary Figures

Figure S1

Moose migration in Fennoscandia

For each moose in our study the centroid of the summer (in green) and winter (in blue) locations are connected with a red line.

Moose migration in Fennoscandia

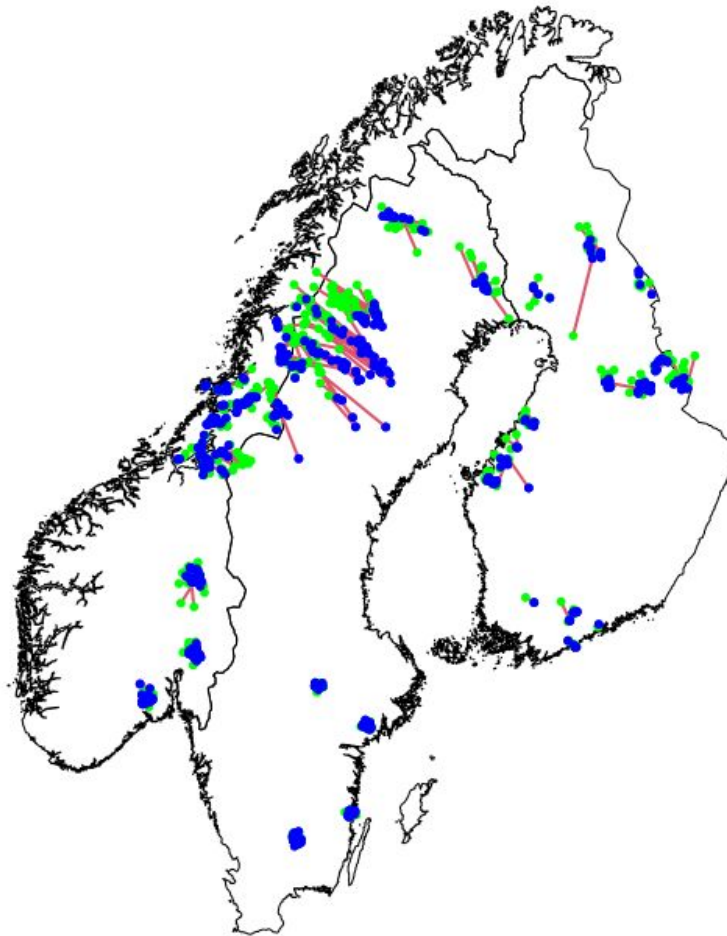


Figure S2

Boxplot with the “leave-one-out” cross validation

The Spearman rank correlation for each moose between the observed and predicted occupancy from a model without the focal moose (i.e. leave-one-out cross validation) for the summer and winter resource selection probability function.

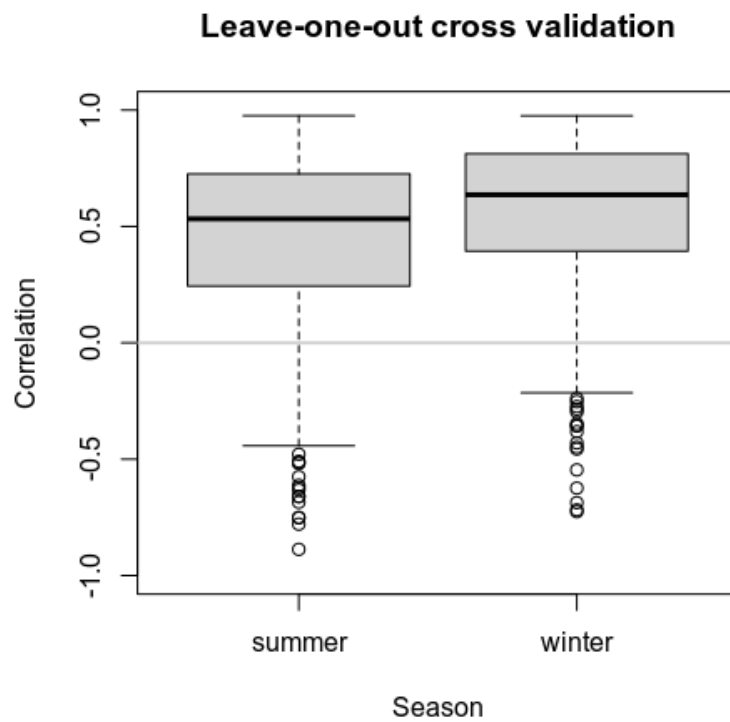


Figure S3

Residuals from the relationship between harvest and habitat suitability in Fennoscandia

The residuals from the linear regression between the average annual harvest and the total summer suitability for each mapping unit in Fennoscandia ($\log(\text{harvest}) = \beta_0 + \beta_1 \log(RSPF_{summer}) + \varepsilon$). The largest deviation from the prediction occurs in the Norwegian west coast, which has extremely low numbers of moose compared with the amount of available habitat. This area is generally not considered part of the moose range in Fennoscandia, however, the mechanisms for this are not well understood. It could be a combination of historical overharvest combined with difficulties in recolonization due to topography and local competition with red deer. If we removed those municipalities from the analysis, we obtained an even stronger relationship between the amount of suitable habitat and moose harvest (see Figure S5). Note that none of the other results, related to moose migration, are affected by the in- or exclusion of the Norwegian west coast, as for obvious reasons we studied migration only in areas where moose are present.

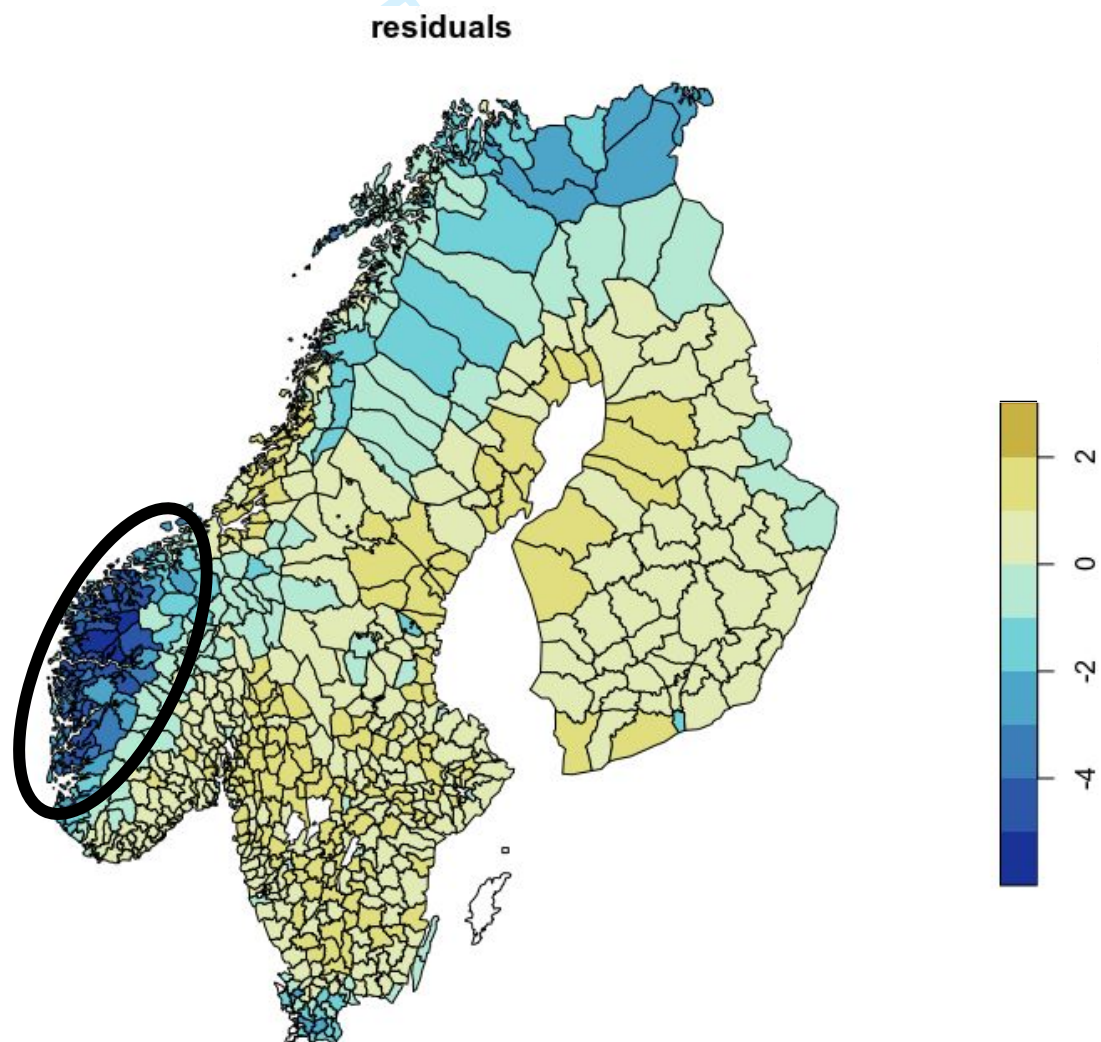


Figure S4

Relationship between harvest and habitat suitability in Fennoscandia

The logarithm of the average annual harvest plotted against the total summer suitability for each reporting unit in Fennoscandia. As noted above, an important deviation from the prediction occurs in Western Norway (see Figure S4). After removal of this area the linear regression ($\log(\text{harvest}) = \beta_0 + \beta_1 \log(\text{RSPF}_{\text{summer}}) + \varepsilon$) explains 66% of the variance in harvest (compared to a variance explained of 50% reported in the main text, and $\beta_1 = 0.86$).

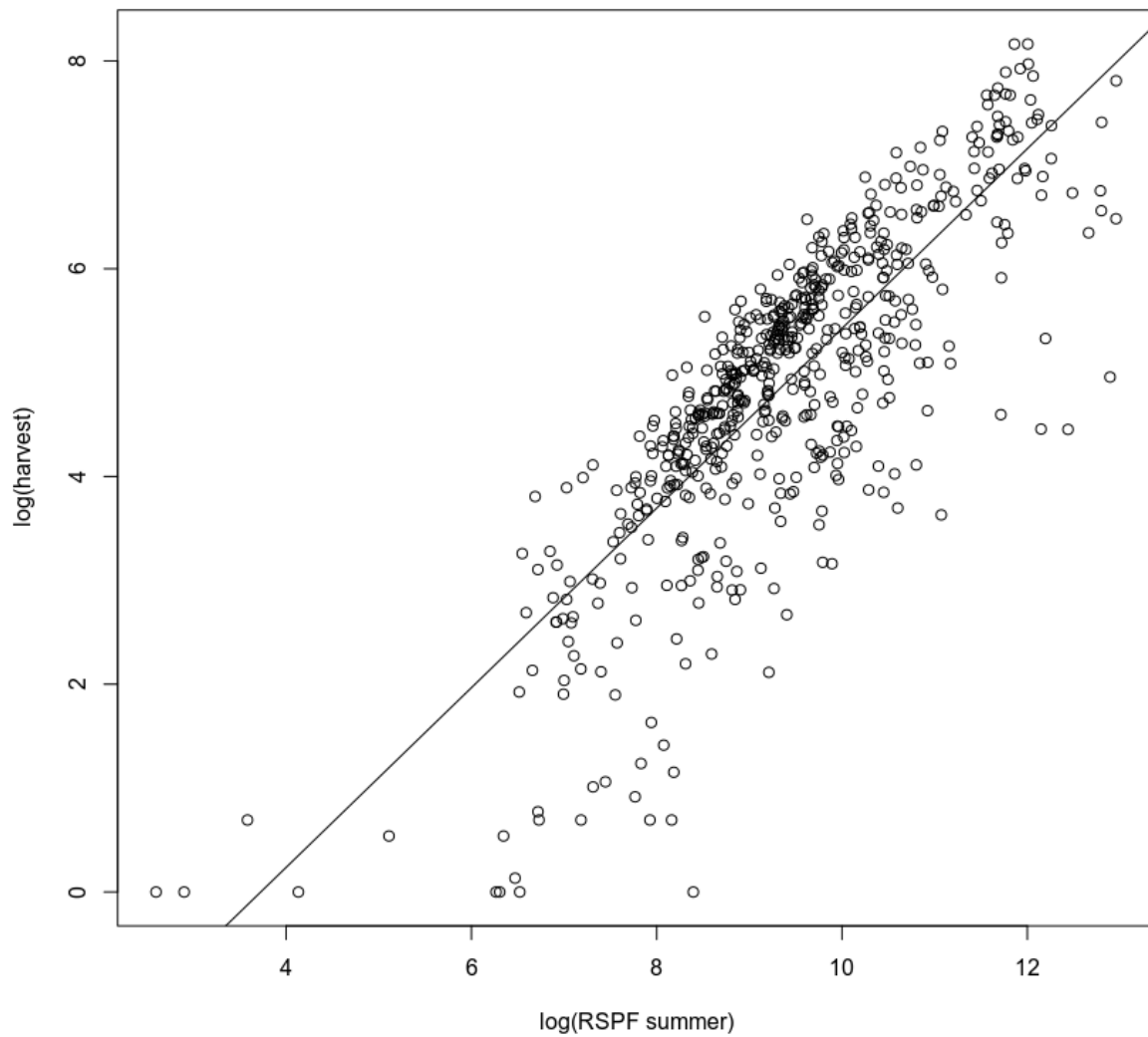


Figure S5

Moose population distribution during summer in Fennoscandia

The population distribution of moose (in harvested individuals per km²) derived from the harvest data and the predicted summer RSPF.

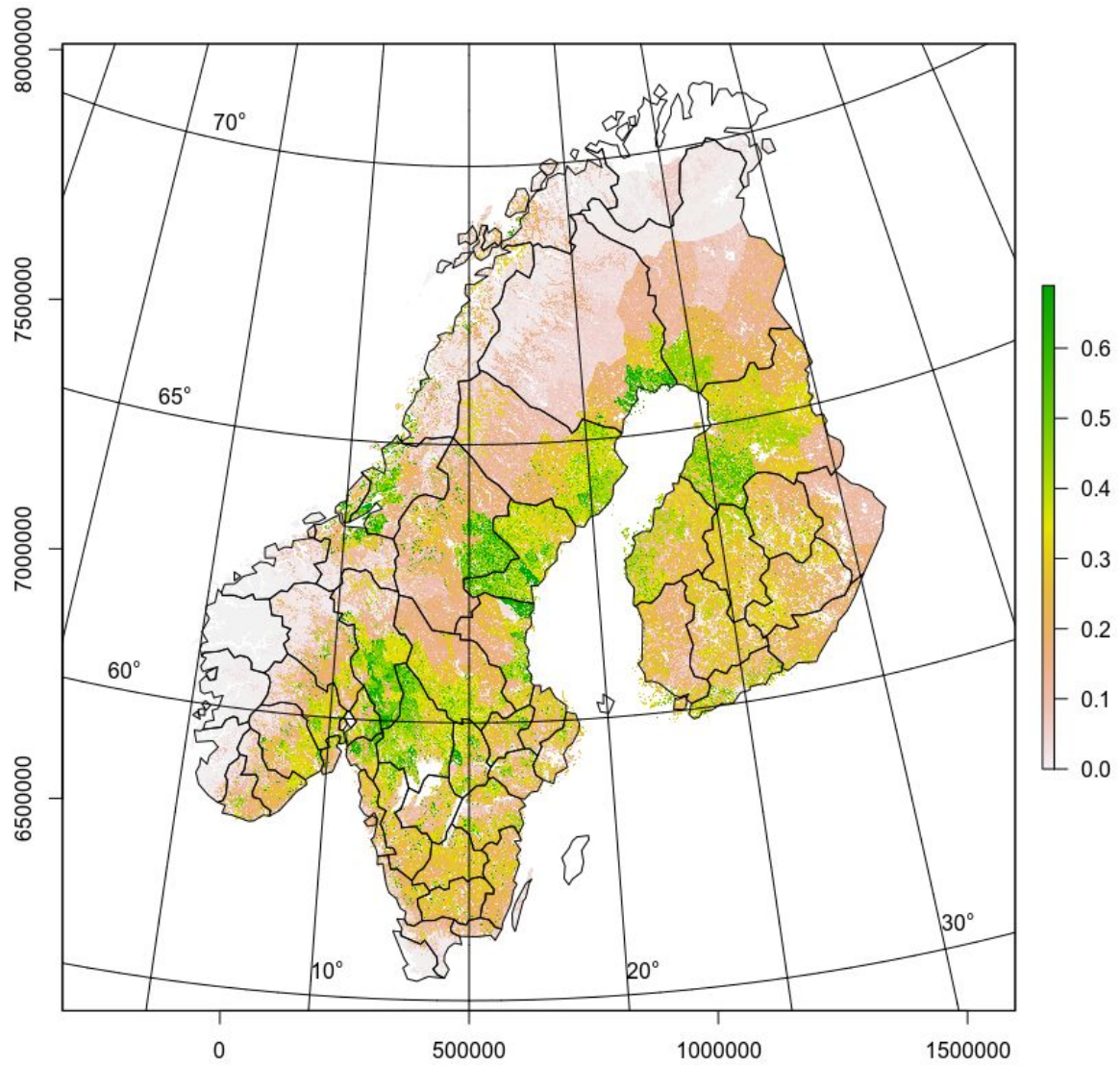


Figure S6

Boxplot with the distribution of elevation at the lowland and highland range across Fennoscandia. We labelled the geographic areas that were more selected during winter than during summer as lowland ranges and those that were more selected during summer than during winter as highland ranges. We used these geographic labels instead of summer and winter ranges to avoid confusion with summer and winter in reference to seasons. Although, altitude is unlikely to be a direct driver of migration itself, it is common to observe altitudinal migration in response to ecological variables with an altitudinal gradient. For instance, in our study snow was an important variable related to migration. Not surprising and in support of our a priori labels, we did observe lower elevations in the areas more selected during winter, and higher elevations in the areas more selected during summer.

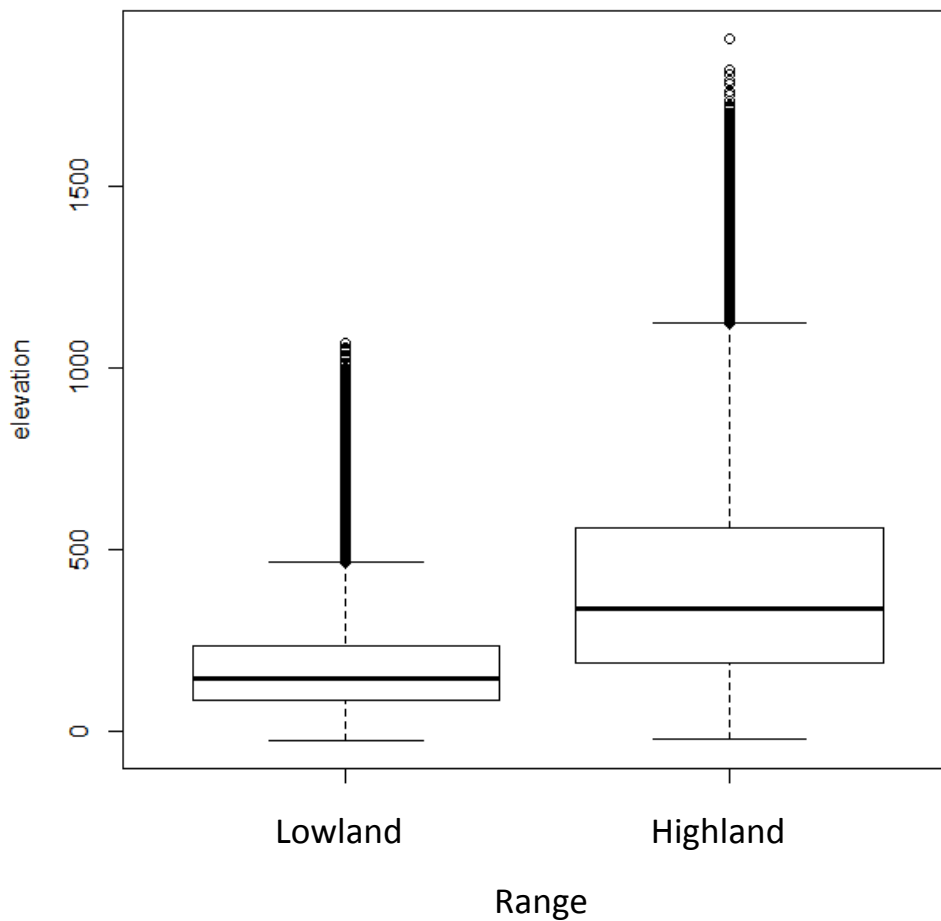


Figure S7

Descriptive details for the different clusters: number of individuals (N), number of males, number of females, number of migrants & residents (VI criterion), mean age of the moose at marking (note that age was not available for all clusters).



Name	N_{total}	N_{male}	N_{female}	$N_{migrant}$	$N_{resident}$	avg(age)
A	30	9	21	29	1	4.9
B	54	8	46	31	23	4.8
C	54	8	46	49	5	5.2
D	19	6	13	16	3	5.5
E	16	4	12	7	9	5.8
F	21	2	19	1	20	8.4
G	42	20	22	31	11	NA
H	29	7	22	15	14	NA
I	13	4	9	5	8	NA
J	13	8	5	7	6	NA
K	24	3	21	0	24	7.1
L	22	5	17	1	21	6.7
M	43	6	37	15	28	3.0
N	31	10	21	17	14	2.7
O	13	2	11	8	5	3.1
P	41	13	28	11	30	NA
Q	22	4	18	2	20	NA
R	26	0	26	13	13	7.1
S	32	0	32	15	17	8.7

Figure S8

Relationship between different migration criteria

The proportion of migrants in each cluster determined by the Volume of Intersection VI criterion on the x-axis and the distance above 10km criterion on the y-axis. Given the high agreement between both criteria results were qualitatively identical, we therefore only reported those for the VI criterion.

