

1 **Effects of single and combined exposures of gold (nano *versus* ionic form)**
2 **and gemfibrozil in a liver organ culture of *Sparus aurata***

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22
23 **Highlights**

- 24
25 • *In vitro*, gold nanoparticles (AuNPs) damaged fish liver DNA and cellular
26 membranes;
27 • Gemfibrozil (GEM) caused DNA damage at 1.5 µg.L⁻¹;

- 28 • Overall, effects of AuNPs+GEM were higher than predicted, based on single
29 exposures;
- 30 • Liver organ culture proved sensitive and a valuable *in vitro* model.

31

32 **Abstract**

33 *In vitro* methods have gained increasing importance in ecotoxicology due to
34 ethical concerns. The aim of this study was to assess the *in vitro* effects of gold, in
35 the nanoparticle (AuNPs) and ionic (Au⁺) form, and the pharmaceutical gemfibrozil
36 (GEM), in single and combined exposures. Fish liver was 24 h exposed to gold (4
37 to 7,200 µg.L⁻¹), GEM (1.5 to 15,000 µg.L⁻¹) and combination 80 µg.L⁻¹ gold + 150
38 µg.L⁻¹ GEM. Endpoints related with antioxidant status, peroxidative and genetic
39 damage were assessed. AuNPs caused more effects than Au⁺, increasing catalase
40 and glutathione reductase activities and damaging DNA and cellular membranes.
41 Effects were dependent on AuNPs size, coating and concentration. GEM damaged
42 DNA at an environmentally relevant concentration, 1.5 µg.L⁻¹. Overall, the effects of
43 the combined exposures were higher than the predicted, based on single
44 exposures. This study showed that liver culture can be a useful model to study
45 contaminants effects.

46

47 **Keywords:** Fish liver culture; gilthead seabream; nanotoxicology; ionic gold;
48 oxidative stress; DNA integrity.

49

50 1. Introduction

51 Due to concerns regarding animal welfare, time and cost constraints, and
52 generation of dangerous residues, establishing workable *in vitro* systems became a
53 priority. In this perspective, the use of organ/cell cultures has the advantage of
54 allowing a reduction in the number of animals used per test, improved control of
55 environmental conditions, reduction of the genetic heterogeneity and chemicals
56 needed, as well as a reduction in waste (Oliveira et al. 2003; Soldatow et al 2013).
57 Liver cell culture models can be important in toxicological research due to the crucial
58 function of this organ in detoxification, metabolic and inflammatory/immune
59 processes (Zeilinger et al. 2016). Although liver cell cultures, including the ones
60 obtained from fish (Franco et al. 2019), are well-established biological
61 methodologies in *in vitro* testing, organ cultures allow the study of effects in a more
62 physiologically relevant context. The use of liver slices retain the 3D structure,
63 contain all liver cell types and show good *in vitro/in vivo* correlation for xenobiotic
64 metabolism (Soldatow et al 2013). Despite the successful use of animals organ
65 cultures in toxicological research, fish organ cultures have not been extensively
66 used in ecotoxicology, despite the reported value to assess the effects of chromium
67 in *Anguilla anguilla* (Oliveira et al. 2003) and to test different oxytetracycline
68 exposure methods, using *Danio rerio* (Chemello et al. 2019).

69 The effects of gold nanoparticles (AuNPs) and the lipid regulator gemfibrozil
70 (GEM) on marine fish remains largely unknown despite their increasing production,
71 use and disposal. Previous *in vivo* studies with gilthead seabream (*Sparus aurata*)
72 have already shown the ability of AuNPs (e.g., Barreto et al. 2020) and GEM (e.g.,
73 Barreto et al. 2017, 2018) to induce toxic effect alone and in combined exposures

74 (Barreto et al. 2019a, 2019b). Considering the reported importance of the liver in
75 AuNPs accumulation (Chen et al. 2013; Iswarya et al. 2016; Khan, Vishakante, and
76 Siddaramaiah 2013; Mateo et al. 2014; Simpson et al. 2013) and metabolization of
77 xenobiotics allied with the need to use methodologies that minimize the need to
78 sacrifice animals providing reliable information in terms of effects and mechanisms
79 of action of emerging contaminants, this study aimed to assess the effects of AuNPs
80 and GEM, alone and in combined exposures, in liver cultures. Effects after 24 h
81 exposure were evaluated measuring endpoints related with oxidative stress,
82 peroxidative and DNA damage.

83

84 **2. Material and Methods**

85 **2.1. Test organisms**

86 Juvenile gilthead seabream (*Sparus aurata*), length 100 ± 0.4 cm, acquired from
87 a Spanish aquaculture facility (Santander, Spain). Fish were acclimated for 4 weeks
88 in 250 L aquaria, at a ratio below 1 g of fish per 1 L of aerated and filtered (Eheim
89 filters) artificial seawater (ASW; Ocean Fish, Prodac), prepared by dissolving the
90 salt in reverse osmosis purified water to obtain a salinity of 30. Fish were maintained
91 in a room-controlled temperature (20 °C) with natural photoperiod. During this
92 period, animals were fed daily with commercial fish food (Sorgal, Portugal) at a ratio
93 of 1% of body weight/day. All experimental procedures followed International
94 Guiding Principles for Biomedical Research Involving Animals (EU 2010/63) and
95 were previously approved by the ethics committee and the responsible national legal
96 authority “Direção Geral de Alimentação e Veterinária” (authorization N421/2013).

97

98 **2.2. Synthesis and characterisation of gold nanoparticles (AuNPs)**

99 Citrate-coated AuNPs (cAuNPs), diameter of 7 nm, were synthesised using the
100 pH-shifting method, with reduction of gold (III) chloride trihydrate by citric acid,
101 followed by neutralization with NaOH (Shiba 2013). cAuNPs, diameter of 40 nm,
102 were prepared, using 15 nm seeds, by sodium citrate reduction of gold (III) chloride
103 trihydrate (Lekeufack et al. 2010). Part of cAuNPs were coated with PVP as
104 described by Barreto et al. (2015). Both coated AuNPs – cAuNPs and PVP coated
105 AuNPs (PVP-AuNPs) – were centrifuged and the pellet resuspended in ultrapure
106 water. AuNPs were characterised in ultrapure water and in the media used for the
107 experiments – Dulbecco's Modified Eagle's medium with fetal bovine serum
108 (DMEM+FBS) – by UV-Vis spectrophotometry (Cintra 303, GBC Scientific) to obtain
109 the UV-Vis spectra; hydrodynamic size was assessed by dynamic light scattering –
110 DLS (Zetasizer Nano ZS, Malvern) and size/shape evaluated by transmission
111 electron microscopy – TEM (Hitachi, H9000 NAR) or scanning electron microscopy
112 – SEM (Hitachi, SU70). Zeta potential (ZP) was measured using Zetasizer (Nano
113 ZS, Malvern). Measurements were performed at 0, 12 and 24 h, at concentrations
114 higher than $80 \mu\text{g.L}^{-1}$, considering that, for concentrations lower than $80 \mu\text{g.L}^{-1}$ the
115 detection limits of the used techniques did not allow the characterisation of AuNPs.
116 The characterisation was also performed visually, assessing the colour of the
117 AuNPs suspensions.

118

119 **2.3. Liver organ culture exposures**

120 DMEM+FBS was prepared as follow: 50% DMEM, 40% ultrapure water, 1 mM of
121 glutamine, 15 mM HEPES, 10% FBS and $100 \mu\text{g.mL}^{-1}$ antibiotics (penicillin and

122 streptomycin). A stock solution of GEM (50 g.L^{-1}) was prepared in dimethyl sulfoxide
123 (DMSO) and test solutions prepared by the dilution of the stock in DMEM+FBS. Test
124 suspensions of AuNPs were prepared in DMEM+FBS from cAuNPs (100 and 97
125 mg.L^{-1} for 7 and 40 nm, respectively) and PVP-AuNPs (78 and 58 mg.L^{-1} for 7 and
126 40 nm, respectively) stock suspensions. Test solutions of Au^+ were prepared by
127 dilution of the stock (2.7 g.L^{-1}) in DMEM+FBS.

128 After the acclimatization period, fish were anesthetized with 100 mg.L^{-1} tricaine
129 methanesulfonate (MS-222) and subsequently euthanized by spinal section. The
130 liver of each animal was carefully removed, washed with fresh phosphate-buffered
131 saline (PBS), cut into small cubes ($2 \times 2 \text{ mm}$) and cultured in an incubator at $25 \text{ }^\circ\text{C}$
132 and 5% CO_2 during 24 h as previously reported (Oliveira et al. 2003). Per fish, six
133 liver cubes per experimental condition were considered. A total of 20 animals was
134 used in this experimental assay corresponding to 5 animals per test repetition (total
135 of tests: 4). The tested concentrations were: 4, 80, 1,600, 3,200, 4,200, 5,200, 6,200
136 and 7,200 $\mu\text{g.L}^{-1}$ of Au (ionic or nano form – 7 and 40 nm; citrate and PVP coating);
137 1.5, 15, 150, 1,500 and 15,000 $\mu\text{g.L}^{-1}$ of GEM and mixture of 80 $\mu\text{g.L}^{-1}$ of Au (ionic
138 or nano form) with 150 $\mu\text{g.L}^{-1}$ GEM. The lowest concentration of AuNPs ($4 \mu\text{g.L}^{-1}$)
139 was selected as a compromise between predicted values of AuNPs for the aquatic
140 environment ($0.14 \mu\text{g.L}^{-1}$) (García-Negrete et al., 2013; Tiede et al., 2009) and the
141 lowest Au concentration detectable limit in the experimental media. The other
142 AuNPs concentrations tested were progressive increases (e.g., 20 or 2-fold
143 increases. Two sizes and two coatings were selected to understand the correlation
144 between the nanoparticles characteristics and their toxic effects. The effects of Au^+
145 were also assessed to allow understanding the nanoparticle specific effect.

146 Concerning GEM, the lowest tested concentration was chosen based on levels
147 detected in surface waters (Fang et al., 2012). The concentration range used to
148 GEM was based on 10-fold increases. The concentrations GEM and Au used for
149 the combined exposures were based on the effects detected in previous *in vivo*
150 studies with *S. aurata* (Barreto et al. 2018, 2019a, 2019b, 2020). A negative control
151 (only DMEM+FBS) and a solvent control with DMSO (at 0.03%, the highest
152 concentration of DMSO used in the GEM treatments) were also performed.
153 Immediately after liver sampling and before organ culture initiation, three liver cubes
154 per fish were stored at -80 °C until further processing. These samples were collected
155 to determine the basal activities/levels of the liver for the assessed endpoints,
156 corresponding to a time 0 h control. Samples of the experimental media were
157 collected at 0 and 24 h for the quantification of Au and GEM. After 24 h exposure,
158 six liver cubes per experimental condition, per fish, were collected: three for
159 biochemical analysis (stored at -80 °C until further processing) and three for DNA
160 integrity assessment (immediately processed).

161

162 **2.4. Quantification of gold and gemfibrozil (GEM) in the experimental media**

163 The determination of Au in the experimental media was performed according to
164 the NIST NCL Method PCC-8 (NIST 2010). After microwave assisted-acid digestion,
165 sample solutions were analyzed by inductively coupled plasma mass spectrometry
166 (ICP-MS) using an iCAP™ Q ICP-MS equipment (Thermo Fisher Scientific).
167 Elemental isotope – ¹⁹⁷Au – was monitored for analytical determination; ¹⁵⁹Tb and
168 ²⁰⁹Bi used as internal standards.

169 The GEM quantification was carried out by solid phase extraction (SPE) as
170 sample preparation technic followed by liquid chromatography (HPLC) with
171 fluorescence detector. SPE was performed using Oasis Strata-X cartridges (200
172 mg, 3 mL) from Phenomenex. Working standard solutions and extracts were
173 analyzed using a Shimadzu LC system equipped with a SIL 20A autosampler, a
174 DGU-20A5 degasser, a LC 20AB pump, and a RF-10AXL fluorescence detector.
175 For the detection and quantification of GEM a Luna column (C18, 5 μ m particle size,
176 4.60x150 mm, Phenomenex) was used. The optimal conditions were found using
177 acetonitrile (eluent B) and 0.1% formic acid in ultrapure water (eluent A), a flow rate
178 of 1.0 mL.min⁻¹, an oven temperature of 30 °C, and an excitation/emission
179 wavelength pair of 210/300 nm. The linear gradient program was run as 5 min from
180 50 to 100% (B) and after 4 min at 100% (B). Injection volume was 40 μ L. More
181 detailed information is presented in the Supplementary Information.

182

183 **2.5. Biochemical analysis**

184 Liver cubes were homogenized in potassium phosphate buffer (0.1 mM; pH 7.4),
185 using an ultrasonic homogenizer (Branson Ultrasonics Sonifier S-250A). The
186 resultant homogenate was divided in two aliquots: one for the evaluation of lipid
187 peroxidation (LPO) levels and the other one for post-mitochondrial supernatant
188 (PMS) isolation. To prevent oxidation, the aliquot of homogenate for LPO levels
189 determination was transferred to a microtube with 4% BHT (2,6-Di-tert-butyl-4-
190 methylphenol) in methanol and stored at -80 °C until analysis. PMS was obtained
191 by centrifugation (12,000 g; 20 min; 4 °C) and aliquots were stored at -80 °C until
192 GST, CAT and GR activities assessment.

193 Protein concentration was determined according to Bradford (1976), adapted to
194 a microplate format, measuring the absorbance at 600 nm and using bovine γ -
195 globuline as a standard.

196 GST activity was determined according to the method of Habig et al. (1974),
197 adapted to a microplate format (Frasco and Guilhermino 2002), following the
198 conjugation of the substrate – 1-chloro-2, 4-dinitrobenzene (CDNB) – with reduced
199 glutathione. Absorbance was recorded at 340 nm and GST activity calculated as
200 nmol CDNB conjugate formed per min per mg of protein ($\epsilon=9.6\times 10^{-3} \text{ M}^{-1}.\text{cm}^{-1}$).

201 CAT activity was assessed according Claiborne (1985). The change in
202 absorbance at 240 nm caused by the dismutation of hydrogen peroxide (H_2O_2) was
203 recorded and CAT activity was evaluated in terms of μmol of H_2O_2 consumed per
204 min per mg of protein ($\epsilon=40 \text{ M}^{-1}.\text{cm}^{-1}$).

205 GR activity was evaluated by the method of Carlberg and Mannervik (1975)
206 adapted to a microplate format (Lima et al. 2007), measuring the reduced
207 nicotinamide-adenine dinucleotide phosphate (NADPH) disappearance at 340 nm
208 and expressed as nmol of oxidized NADPH (NADP^+) produced per min per mg of
209 protein ($\epsilon=6.22\times 10^3 \text{ M}^{-1}.\text{cm}^{-1}$).

210 LPO levels were measured by the formation of thiobarbituric acid reactive
211 substances (TBARS) based on Ohkawa et al. (1979), adapted by Filho et al. (2001).
212 Absorbance was evaluated at 535 nm and LPO levels expressed as nmol of TBARS
213 produced per mg of protein ($\epsilon=1.56\times 10^5 \text{ M}^{-1}.\text{cm}^{-1}$).

214

215 **2.6. DNA integrity assessment**

216 The alkaline comet assay was performed according to method of Singh et al.
217 (1988) with some adaptations. Each liver cube was disrupted in PBS (pH 7.4) to
218 obtain a suspension. This suspension was centrifuged, the supernatant was
219 discarded, and the pellet was resuspended in fresh PBS. Then, cell suspension was
220 added to 1% (w/v) low melting point agarose (at 40 °C) and the mixture added to a
221 microscope slide pre-coated with 1% (w/v) of normal melting point agarose.
222 Solidification of agarose was allowed by keeping the slides on ice for 5 min. Positive
223 controls (cell suspensions treated with 25 µM of H₂O₂ during 10 min) were included
224 for each electrophoresis run to verify that the electrophoresis conditions were
225 adequate. To lyse the cells, the slides were subsequently immersed in prepared ice-
226 cold lysis solution (2.5 M NaCl, 100 mM EDTA and 10 mM Tris; pH 10.0) containing
227 freshly added 1% Triton X-100 for 1 h, at 4 °C, in the dark. The slides were incubated
228 in alkaline buffer (300 mM NaOH and 1 mM EDTA; pH>13) during 20 min for DNA
229 unwinding. Electrophoresis was performed in the same buffer for 30 min by applying
230 an electric field of 20 V and adjusting the current to 300 mA. After the
231 electrophoresis, the slides were washed with 400 mM Tris-HCl buffer (pH 7.5). The
232 slides were also dehydrated with absolute ethanol and left to dry in the dark. Slides
233 were stained with ethidium bromide (20 µL.mL⁻¹), covered with a coverslip and then
234 visualised using a fluorescence microscope (Olympus BX41TF) at 400X
235 magnification.

236 Slides were analysed randomly, by counting one hundred cells per slide,
237 arbitrarily selected. Cells were scored visually, according to tail length, into 5 classes
238 – from class 0 to 4 (Collins 2004). A damage index (DI) expressed in arbitrary units

239 was assigned to each replicate and consequently for each treatment, according to
240 the damage classes, applying the formula:

241

$$242 \quad DI = (0 \times n_0) + (1 \times n_1) + (2 \times n_2) + (3 \times n_3) + (4 \times n_4)$$

243 where n = number of cells in each class analysed. DI can vary from 0 to 400.

244

245 **2.7. Data analysis**

246 First, Shapiro-Wilk and Levene's tests were used to assess the normality and
247 homogeneity of variance of the data, using the Sigma Plot 12.0 software package.

248 Differences between controls (negative and solvent) were examined using a Student
249 t-test. To detect significant differences between the control and AuNPs single
250 treatments, a two-way analysis of variance (ANOVA) was performed, using
251 concentration and coating as factors, followed by a Dunnett's test. In addition,
252 differences between Au⁺, GEM, the mixtures and control were tested using a one-
253 way ANOVA, followed by Dunnett's test. One-way ANOVA, followed by Tukey's test,
254 whenever applicable, was used to compare differences between AuNPs, Au⁺ and
255 GEM single treatments. Significant differences were accepted for $p < 0.05$.

256 In the combined exposures, the observed effects (in percentage) were compared
257 with the predicted effects (in percentage) obtained by the sum of single exposure
258 effects. This analysis was performed to understand if the combined effect of Au
259 (nano or ionic form) and GEM was lower, similar or greater than the sum of single
260 exposure effects.

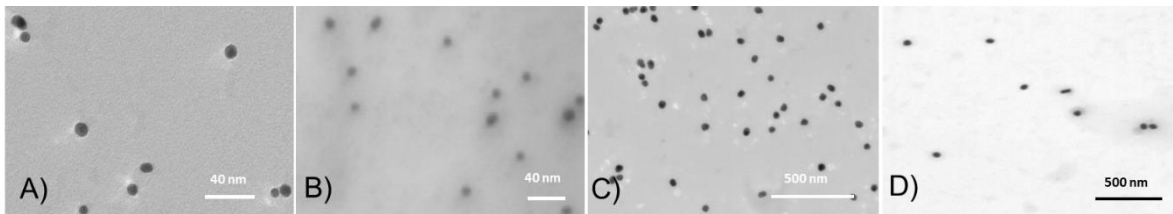
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262 **3. Results**

263 **3.1. Characterisation and behaviour of gold nanoparticles (AuNPs)**

264 The synthesized cAuNPs displayed a round shape (Figure 1A and C), a well-
265 defined absorption band and negative surface charge (Table S1) associated with
266 the citrate layer. The analysis of the size, taking into account the results obtained by
267 DLS and TEM images, revealed an expected average size around 7 and 40 nm.
268 PVP coating led to an increased size due to a PVP layer, detectable by SEM in
269 some AuNPs (Figure 1B and D). The UV-Vis spectra revealed a slight shift in surface
270 plasmon resonance (SPR) peak to longer wavelength when compared with the
271 original cAuNPs (Table S1). ZP shifted from -43 to around -13 mV and from -44 to -
272 17 mV, for 7 and 40 nm AuNPs, respectively (Table S1).

273



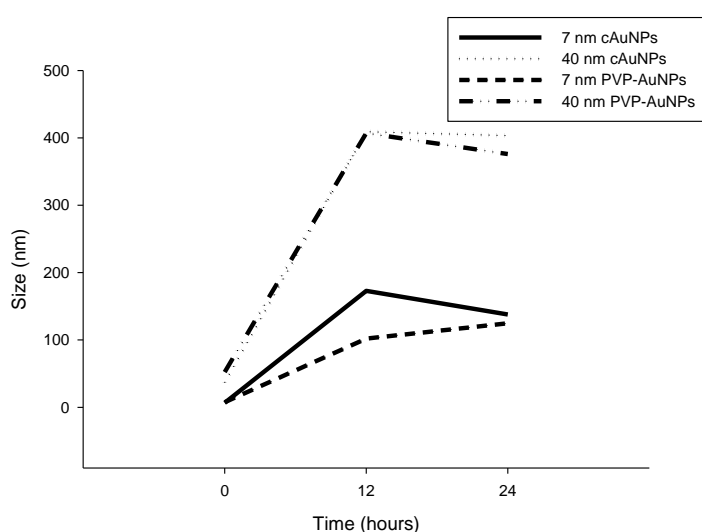
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275 **Figure 1.** Electron microscopy images of 7 and 40 nm citrate (cAuNPs) and
276 polyvinylpyrrolidone (PVP-AuNPs) gold nanoparticles stock suspensions in
277 ultrapure water: **A)** 7 nm cAuNPs (100 mg.L⁻¹); **B)** 7 nm PVP-AuNPs (78 mg.L⁻¹); **C)**
278 40 nm cAuNPs (97 mg.L⁻¹); **D)** 40 nm PVP-AuNPs (58 mg.L⁻¹).

279

280 In ultrapure water, AuNPs were stable, with no detectable
281 agglomerates/aggregates (Figure 1). Size, ZP and UV-Vis spectra of each type of
282 AuNPs were similar during the assessed periods 0, 12 and 24 h. In DMEM+FBS, at
283 0 h, the characteristics of each type of AuNPs were similar to those in ultrapure
284 water, with a slight less negative ZP, slight 1-4 nm increased sizes and shifted SPR

285 peaks toward higher wavelengths (increased about 2–4 nm). Within 12 h, for
286 concentrations higher than $1,600 \mu\text{g}\cdot\text{L}^{-1}$, AuNPs aggregated/agglomerated, with
287 sizes, assessed by DLS, bigger than 100 nm (Figure 2) and SPR peaks shifted
288 toward higher wavelengths (Figure S1). Alterations in ZP were also found, with
289 different peaks correspondent to different charges. After 24 h, no alterations in the
290 size were found, comparing with 12 h (Figure 2) but the SPR peak disappeared
291 (Figure S1).



292

293 **Figure 2.** Size of gold nanoparticles (AuNPs), measured by dynamic light
294 scattering (DLS), in Dulbecco's Modified Eagle's medium with fetal bovine serum
295 (DMEM+FBS) at 0, 12 and 24 h. cAuNPs – Citrate coated gold nanoparticles; PVP-
296 AuNPs – Polyvinylpyrrolidone coated gold nanoparticles.

297

298 The colour of the AuNPs in DMEM+FBS, at 12 h, was between red and blue,
299 being bluer in the highest concentrations. At 24 h, some dark sediment was found
300 in the bottom of the wells. This sediment increased with the increase of AuNPs
301 concentration. At the lower tested concentrations (4 and $80 \mu\text{g}\cdot\text{L}^{-1}$), the media did
302 not present the typical colour of AuNPs agglomeration/aggregation.

303 **3.2. Quantification of gold and gemfibrozil (GEM) in the experimental media**

304 At 0 h, in general, the amount of Au quantified in the experimental media
305 (DMEM+FBS) was lower than the nominal concentrations. After 24 h of exposure,
306 the concentration of Au decreased, particularly after exposure to AuNPs (Table S2).
307 Concerning GEM, at 0 h, measured concentrations were lower than the nominal
308 concentrations, with exception of the concentrations 1.5 and 15 $\mu\text{g.L}^{-1}$ – Table S2.
309 After 24 h, the concentration of GEM decreased slightly.

310

311 **3.3. Biological effects**

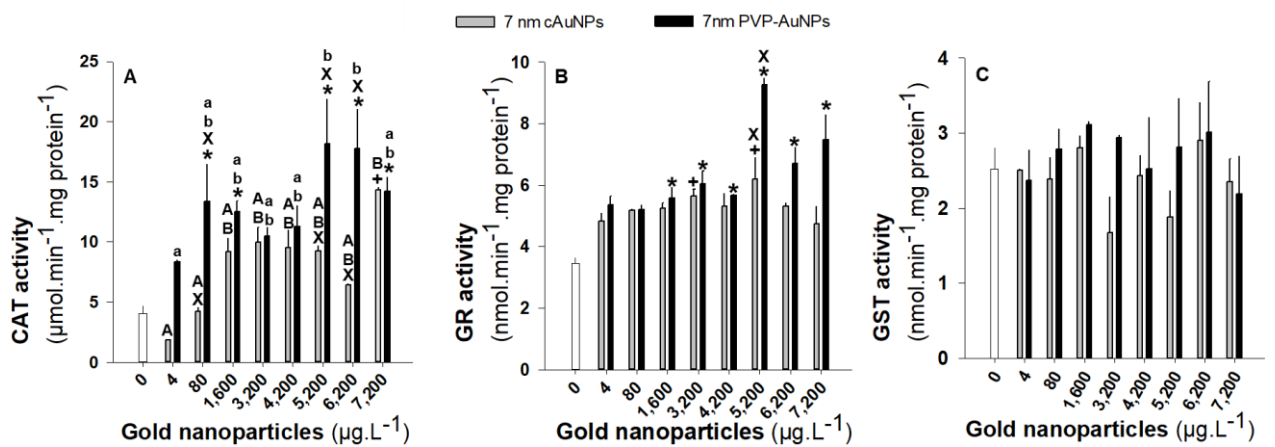
312 For all the tested endpoints, no significant differences ($p>0.05$; t-test) were found
313 between the samples collected immediately after liver sampling and before organ
314 culture initiation (control at 0 h) and controls in DMEM+FBS after 24 h culture.
315 Moreover, after 24 h liver organ culture, no significant differences were found
316 between control and solvent control groups in terms of the tested endpoints ($p>0.05$;
317 t-test). Therefore, the treatments were compared to the control.

318

319 **3.3.1. Effects of 7 nm gold nanoparticles (AuNPs)**

320 For the smallest tested AuNPs, coating promoted different response patterns.
321 The cAuNPs only affected CAT activity at the highest concentration, increasing it
322 ($p<0.05$; Dunnett's test), whereas almost all concentrations tested of PVP-AuNPs
323 (except 4, 3,200 and 4,200 $\mu\text{g.L}^{-1}$) increased CAT activity ($p<0.05$; Dunnett's test;
324 Figure 3A). At 80, 5,200 and 6,200 $\mu\text{g.L}^{-1}$, PVP-AuNPs increased significantly more
325 the CAT activity than cAuNPs ($p<0.05$; Dunnett's test; Figure 3A). Also, the highest
326 concentrations (5,200, 6,200 and 7,200 $\mu\text{g.L}^{-1}$) increased significantly more the CAT

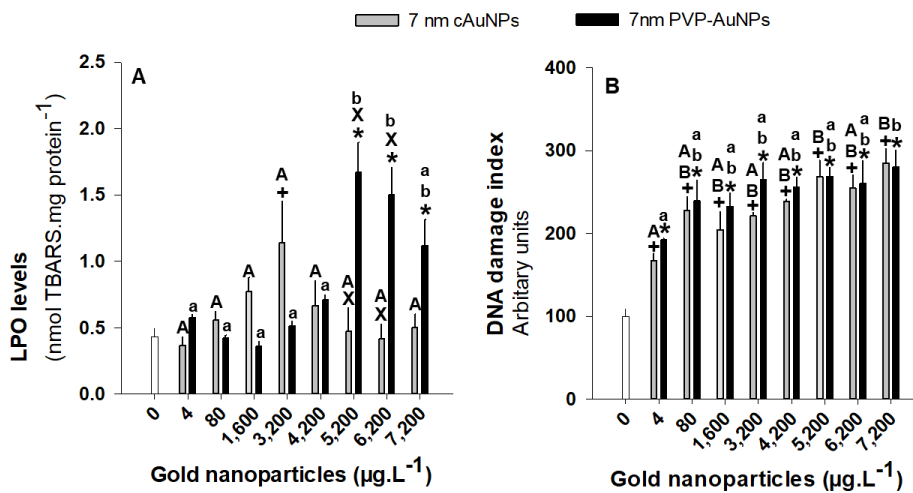
327 activity than the lowest concentrations (4 and 80 $\mu\text{g.L}^{-1}$) ($p < 0.05$; Tukey's test;
 328 Figure 3A). In terms of GR, activity was increased after exposure to 3,200 and 5,200
 329 $\mu\text{g.L}^{-1}$ cAuNPs and to PVP-AuNPs, at concentrations higher than 80 $\mu\text{g.L}^{-1}$ ($p < 0.05$;
 330 Dunnett's test; Figure 3B). At 5,200 $\mu\text{g.L}^{-1}$, PVP-AuNPs increased significantly more
 331 the GR activity than cAuNPs ($p < 0.05$; Dunnett's test; Figure 3A). GST activity was
 332 not significantly affected by exposure to 7 nm AuNPs ($p > 0.05$; ANOVA; Figure 3C).



333

334 **Figure 3.** Catalase (CAT) **(A)**, glutathione reductase (GR) **(B)** and glutathione S-
 335 transferases (GST) **(C)** activities in *Sparus aurata* liver after 24 h organ culture
 336 exposure to 7 nm gold nanoparticles. Results are expressed as mean \pm standard
 337 error. +Significant differences to control (Dunnett's test, $p < 0.05$, citrate coating).
 338 *Significant differences to control (Dunnett's test, $p < 0.05$, polyvinylpyrrolidone
 339 coating). XSignificant differences between cAuNPs and PVP-AuNPs within the same
 340 concentration (Dunnett's test, $p < 0.05$). Different letters correspond to significant
 341 differences between the concentrations of each type of AuNPs, capital letters to
 342 cAuNPs and small letters to PVP-AuNPs (Tukey's test, $p < 0.05$). Citrate coated gold
 343 nanoparticles – cAuNPs; Polyvinylpyrrolidone coated gold nanoparticles – PVP-
 344 AuNPs.

345 The 7 nm AuNPs displayed ability to induce peroxidative damage in membranes.
 346 This effect was more consistent for PVP-AuNPs, that induced increased TBARS at
 347 concentrations higher than 4,200 $\mu\text{g.L}^{-1}$ whereas for cAuNPs, effects were only
 348 found at 3,200 $\mu\text{g.L}^{-1}$ ($p < 0.05$; Dunnett's test; Figure 4A). Indeed, at 5,200 and 6,200
 349 $\mu\text{g.L}^{-1}$, PVP-AuNPs induced significantly higher levels of peroxidative damage than
 350 cAuNPs ($p < 0.05$; Dunnett's test; Figure 4A). At 5,200 and 6,200 $\mu\text{g.L}^{-1}$, PVP-AuNPs
 351 induced significantly higher levels of LPO than at 4 to 4,200 $\mu\text{g.L}^{-1}$ ($p < 0.05$; Tukey's
 352 test; Figure 4A). All the 7 nm AuNPs tested concentrations induced DNA damage
 353 ($p < 0.05$; Dunnett's test; Figure 4B), with 5,200 and 7,200 $\mu\text{g.L}^{-1}$ inducing significantly
 354 more DNA damage than 4 $\mu\text{g.L}^{-1}$ ($p < 0.05$; Tukey's test; Figure 4B).



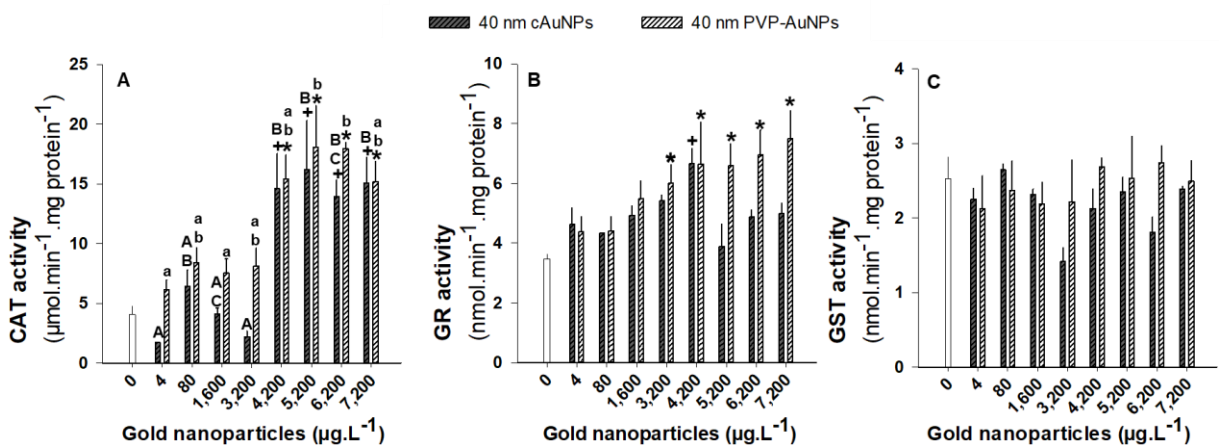
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 356 **Figure 4.** Lipid peroxidation (LPO) levels (A) and DNA damage index (B) of
 357 *Sparus aurata* liver after 24 h organ culture exposure to 7 nm gold nanoparticles.
 358 Results are expressed as mean \pm standard error. +Significant differences to control
 359 (Dunnett's test, $p < 0.05$, citrate coating). *Significant differences to control (Dunnett's
 360 test, $p < 0.05$, polyvinylpyrrolidone coating). XSignificant differences between
 361 cAuNPs and PVP-AuNPs within the same concentration (Dunnett's test, $p < 0.05$).

362 Different letters correspond to significant differences between the concentrations of
 363 each type of AuNPs, capital letters to cAuNPs and small letters to PVP-AuNPs
 364 (Tukey's test, $p < 0.05$). Citrate coated gold nanoparticles – cAuNPs;
 365 Polyvinylpyrrolidone coated gold nanoparticles – PVP-AuNPs.

366

367 3.3.2. Effects of 40 nm gold nanoparticles (AuNPs)

368 Effects of 40 nm AuNPs on CAT activity were found, for both coatings, at
 369 concentrations higher than 3,200 $\mu\text{g.L}^{-1}$ ($p < 0.05$; Dunnett's test; Figure 5A). The
 370 highest tested concentrations (4,200 to 7,200 $\mu\text{g.L}^{-1}$) increased significantly more
 371 the CAT activity than the lowest tested concentrations (4 to 3,200 $\mu\text{g.L}^{-1}$) ($p < 0.05$;
 372 Tukey's test; Figure 5A). These particles also induced an increase in GR activity at
 373 4,200 $\mu\text{g.L}^{-1}$ for cAuNPs and concentrations higher than 1,600 $\mu\text{g.L}^{-1}$ for PVP-
 374 AuNPs ($p < 0.05$; Dunnett's test; Figure 5B). As observed in liver culture exposed to
 375 7 nm AuNPs, GST activity was not significantly affected by 40 nm AuNPs ($p > 0.05$;
 376 ANOVA; Figure 5C).

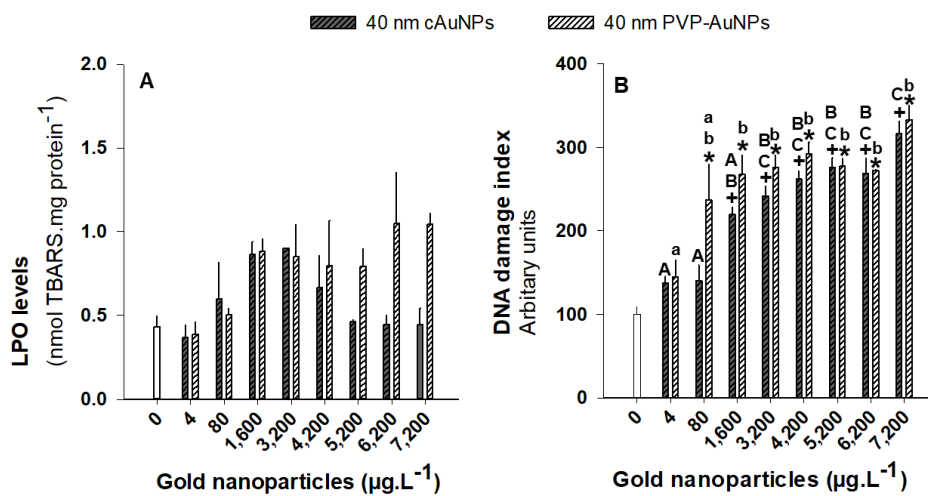


377

378 **Figure 5.** Catalase (CAT) (A), glutathione reductase (GR) (B) and glutathione S-
 379 transferases (GST) (C) activities of *Sparus aurata* liver after 24 h organ culture

380 exposure to 40 nm gold nanoparticles. Results are expressed as mean \pm standard
 381 error. *Significant differences to control (Dunnett's test, $p < 0.05$, citrate coating).
 382 *Significant differences to control (Dunnett's test, $p < 0.05$, polyvinylpyrrolidone
 383 coating). Different letters correspond to significant differences between the
 384 concentrations of each type of AuNPs, capital letters to cAuNPs and small letters to
 385 PVP-AuNPs (Tukey's test, $p < 0.05$). Citrate coated gold nanoparticles – cAuNPs;
 386 Polyvinylpyrrolidone coated gold nanoparticles – PVP-AuNPs.

387
 388 No significant oxidative damage, assessed as LPO, was found after liver
 389 exposure to 40 nm AuNPs ($p > 0.05$; ANOVA; Figure 6A). However, DNA damage
 390 was found after exposure to concentrations higher than 4 $\mu\text{g.L}^{-1}$ for PVP-AuNPs and
 391 80 $\mu\text{g.L}^{-1}$ for cAuNPs ($p < 0.05$; Dunnett's test; Figure 6B). The highest tested
 392 concentrations (e.g., 3,200 to 7,200 $\mu\text{g.L}^{-1}$ cAuNPs) induced significantly more DNA
 393 damage when compared with the lowest tested concentrations (e.g., 4 and 80 $\mu\text{g.L}^{-1}$
 394 cAuNPs) ($p < 0.05$; Tukey's test; Figure 6B).

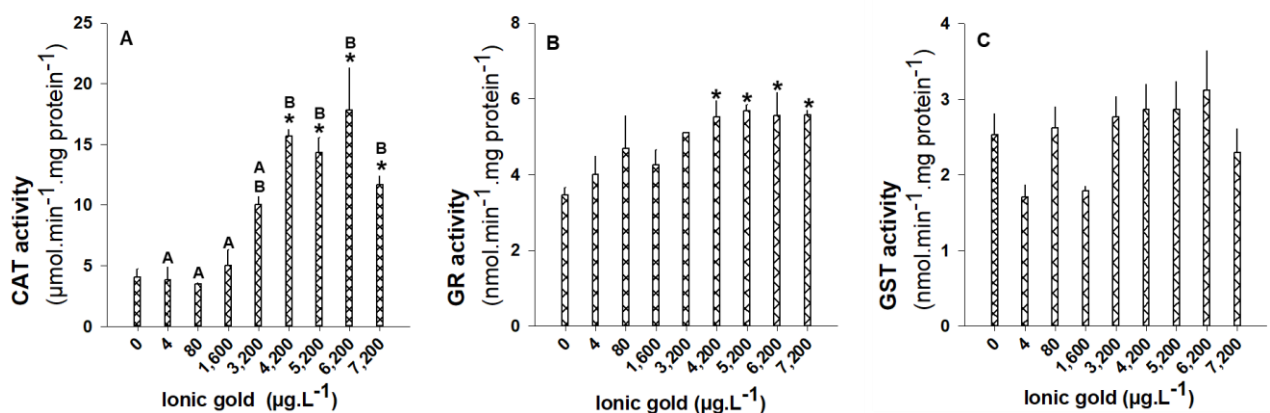


395

396 **Figure 6.** Lipid peroxidation (LPO) levels **(A)** and DNA damage index **(B)** of
 397 *Sparus aurata* liver after 24 h organ culture exposure to 40 nm gold nanoparticles.
 398 Results are expressed as mean \pm standard error. ⁺Significant differences to control
 399 (Dunnett's test, $p < 0.05$, citrate coating). *Significant differences to control (Dunnett's
 400 test, $p < 0.05$, polyvinylpyrrolidone coating). Different letters correspond to significant
 401 differences between the concentrations of each type of AuNPs, capital letters to
 402 cAuNPs and small letters to PVP-AuNPs (Tukey's test, $p < 0.05$). Citrate coated gold
 403 nanoparticles – cAuNPs; Polyvinylpyrrolidone coated gold nanoparticles – PVP-
 404 AuNPs.

406 3.3.3. Effects of ionic gold (Au⁺)

407 At concentrations higher than 3,200 $\mu\text{g.L}^{-1}$, Au⁺ significantly increased the
 408 activities of CAT and GR ($p < 0.05$; Dunnett's test; Figure 7A and B). The highest
 409 tested concentrations (4,200 to 7,200 $\mu\text{g.L}^{-1}$) increased significantly more the CAT
 410 activity than the lowest tested concentrations (4 to 1,600 $\mu\text{g.L}^{-1}$) ($p < 0.05$; Tukey's
 411 test; Figure 7A). As observed for AuNPs, liver exposure to Au⁺ did not induce
 412 significant alterations in GST activity ($p > 0.05$; ANOVA; Figure 7C).



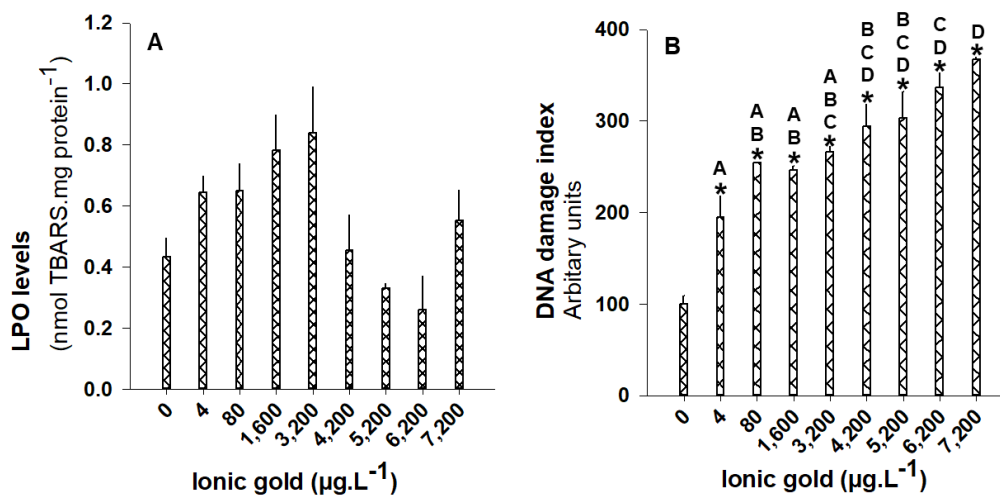
413

414 **Figure 7.** Catalase (CAT) **(A)**, glutathione reductase (GR) **(B)** and glutathione S-
 415 transferases (GST) **(C)** activities of *Sparus aurata* liver after 24 h organ culture
 416 exposure to ionic gold. Results are expressed as mean \pm standard error. *Significant
 417 differences to control (Dunnett's test, $p < 0.05$). Different letters correspond to
 418 significant differences between the concentrations (Tukey's test, $p < 0.05$).

419

420 The LPO levels remained unchanged after the exposure to Au^+ ($p > 0.05$; ANOVA;
 421 Figure 8A). However, DNA damage was found after exposure to all the tested
 422 concentrations ($p < 0.05$; Dunnett's test; Figure 8B). The highest tested
 423 concentrations (e.g., 6,200 and 7,200 $\mu\text{g.L}^{-1}$) induced significantly more DNA
 424 damage when compared with the lowest tested concentrations (e.g., 4 to 1600 $\mu\text{g.L}^{-1}$
 425 $^{137}\text{CsAuNPs}$) ($p < 0.05$; Tukey's test; Figure 8B).

426



427

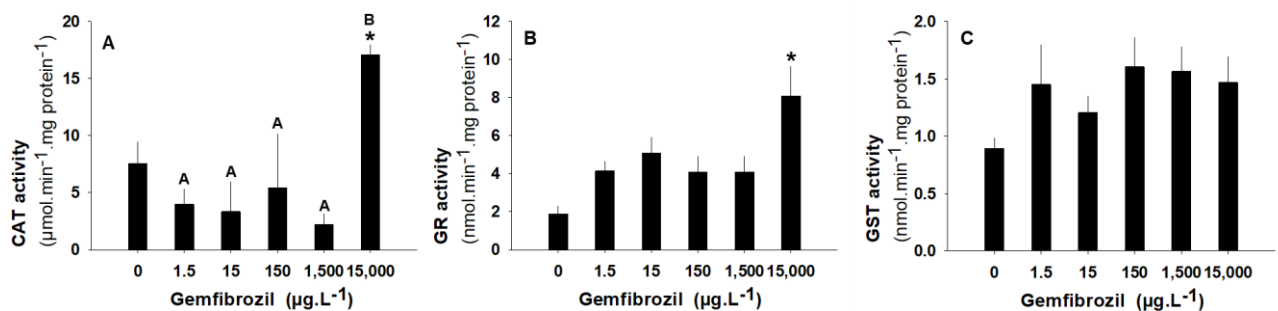
428 **Figure 8.** Lipid peroxidation (LPO) levels **(A)** and DNA damage index **(B)** of
 429 *Sparus aurata* liver after 24 h organ culture exposure to ionic gold. Results are
 430 expressed as mean \pm standard error. *Significant differences to control (Dunnett's

431 test, $p < 0.05$). Different letters correspond to significant differences between the
432 concentrations (Tukey's test, $p < 0.05$).

433

434 3.3.4. Effects of gemfibrozil (GEM)

435 CAT and GR activities were significantly increased after exposure to 15,000 $\mu\text{g.L}^{-1}$
436 1 GEM ($p < 0.05$; Dunnett's test; Figure 9A and B). GEM, at 15,000 $\mu\text{g.L}^{-1}$, significantly
437 increased CAT activity comparing with the other tested concentrations (1.5 to 1,500
438 $\mu\text{g.L}^{-1}$) ($p < 0.05$; Tukey's test; Figure 9A). However, GST activity was not significantly
439 affected by exposure to GEM ($p > 0.05$; ANOVA; Figure 9C).



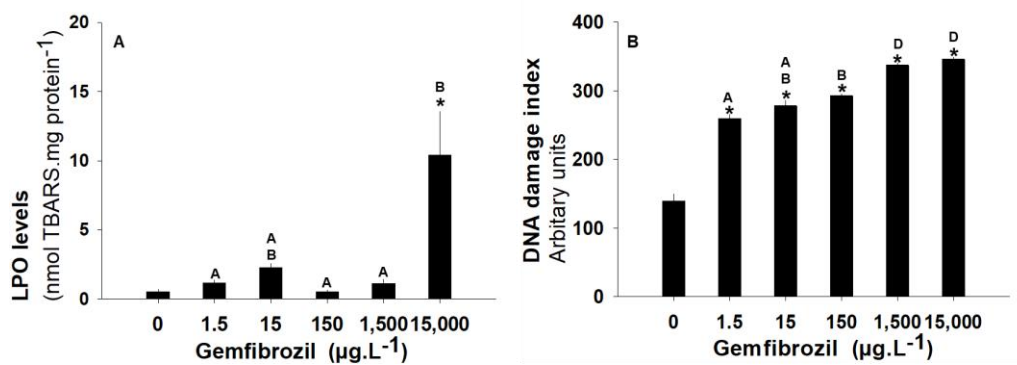
440

441 **Figure 9.** Catalase (CAT) **(A)**, glutathione reductase (GR) **(B)** and glutathione S-
442 transferases (GST) **(C)** activities of *Sparus aurata* liver after 24 h organ culture
443 exposure to gemfibrozil. Results are expressed as mean \pm standard error.
444 *Significant differences to control (Dunnett's test, $p < 0.05$). Different letters
445 correspond to significant differences between the concentrations (Tukey's test,
446 $p < 0.05$).

447

448 LPO levels significantly increased after exposure to 15,000 $\mu\text{g.L}^{-1}$ GEM ($p < 0.05$;
449 Dunnett's test; Figure 10A). GEM, at 15,000 $\mu\text{g.L}^{-1}$, induced significantly higher
450 levels of LPO than 1.5, 150 and 1,500 $\mu\text{g.L}^{-1}$ GEM ($p < 0.05$; Tukey's test; Figure

451 10A). In terms of DNA damage, all tested GEM concentrations led to a significant
 452 decrease in the DNA integrity ($p < 0.05$; Dunnett's test; Figure 10B). The highest
 453 tested concentrations, 1,500 and 15,000 $\mu\text{g.L}^{-1}$, induced significantly more DNA
 454 damage than the lowest tested concentrations (1.5 to 150 $\mu\text{g.L}^{-1}$ cAuNPs) ($p < 0.05$;
 455 Tukey's test; Figure 10B).



456

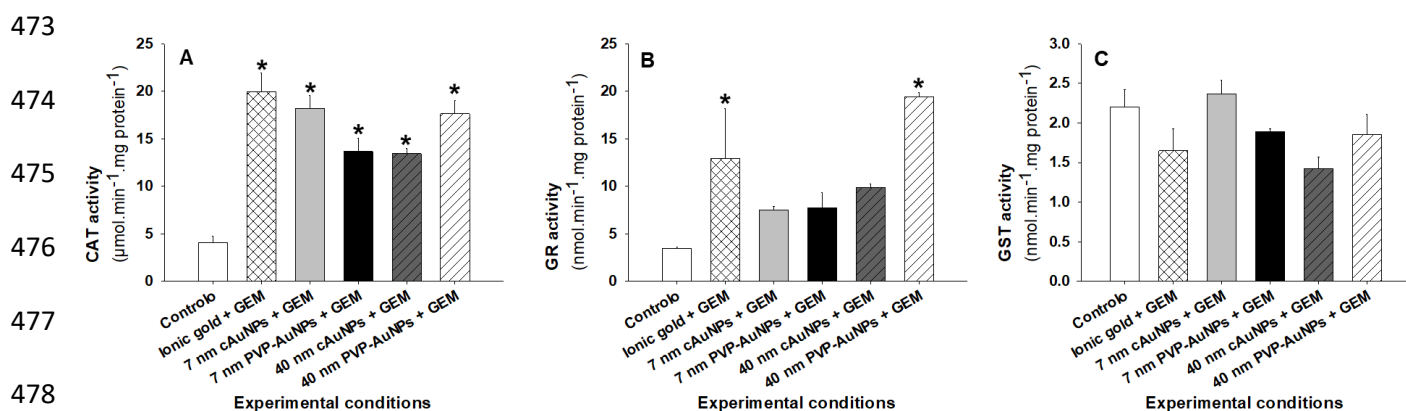
457 **Figure 10.** Lipid peroxidation (LPO) levels (A) and DNA damage index (B) of
 458 *Sparus aurata* liver after 24 h organ culture exposure to gemfibrozil. Results are
 459 expressed as mean \pm standard error. *Significant differences to control (Tukey's test,
 460 $p < 0.05$). Different letters correspond to significant differences between the
 461 concentrations (Tukey's test, $p < 0.05$).

462

463 3.3.5. Effects of combined exposures: gold and gemfibrozil (GEM)

464 In the combined exposures, CAT activity significantly increased ($p < 0.05$;
 465 Dunnett's test; Figure 11A), with observed percentages of effect being higher than
 466 the predicted based on the single exposures (Table 1). The combined exposures to
 467 Au⁺ + GEM and 40 nm PVP-AuNPs + GEM significantly increased GR activity
 468 ($p < 0.05$; Dunnett's test; Figure 11B), with observed percentages of effect being
 469 higher than the predicted (Table 1). GST activity was not significantly affected by
 470 the combined exposures ($p > 0.05$; ANOVA; Figure 11C), as observed in the single

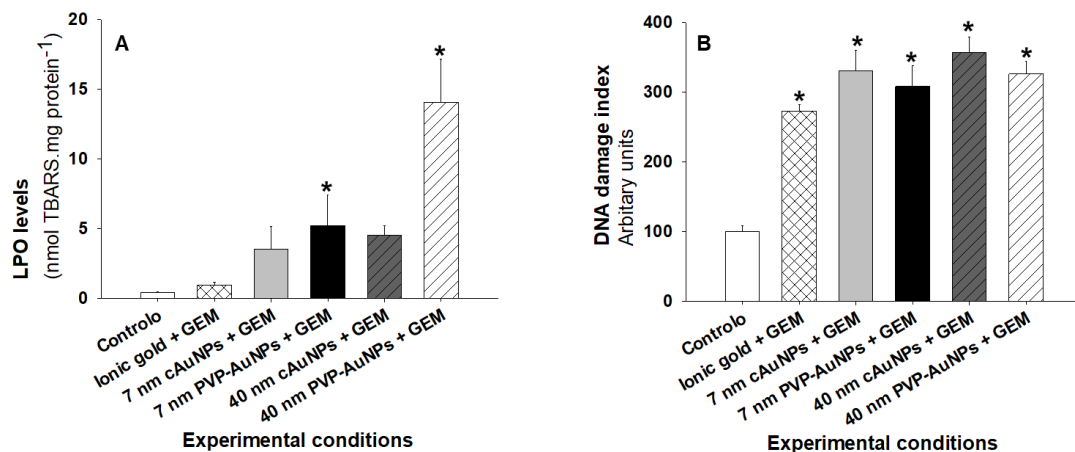
471 exposures. For this endpoint, the observed percentages of effect were the predicted
 472 (Table 1).



479 **Figure 11.** Catalase (CAT) **(A)** glutathione reductase (GR) **(B)** and glutathione S-
 480 transferases (GST) **(C)** activities of *Sparus aurata* liver after 24 h organ culture
 481 combined exposure to gold nanoparticles (AuNPs) or ionic gold (Au⁺) with
 482 gemfibrozil (GEM). Results are expressed as mean ± standard error. *Significant
 483 differences to control (Dunnett’s test, $p < 0.05$). Citrate coated gold nanoparticles –
 484 cAuNPs; Polyvinylpyrrolidone coated gold nanoparticles – PVP-AuNPs; Au + GEM
 485 – 80 µg.L⁻¹ Au⁺ or AuNPs (cAuNPs or PVP-AuNPs) with 150 µg.L⁻¹ GEM.

486

487 The combined exposures to PVP-AuNPs (7 and 40 nm) and GEM significantly
 488 increased LPO levels ($p < 0.05$; Dunnett’s test; Figure 12A), yielding observed
 489 percentages of effect higher than the predicted (Table 1). All the combined
 490 exposures induced significant increases in DNA damage ($p < 0.05$; Dunnett’s test;
 491 Figure 12B), with the observed percentages of effect similar to those expected
 492 (Table 1).



493

494 **Figure 12.** Lipid peroxidation (LPO) levels **(A)** and DNA damage index (arbitrary
 495 units) **(B)** of *Sparus aurata* liver after 24 h organ culture combined exposure to gold
 496 nanoparticles (AuNPs) or ionic gold (Au⁺) with gemfibrozil (GEM). Results are
 497 expressed as mean ± standard error. *Significant differences to control (Dunnett's
 498 test, $p < 0.05$). Citrate coated gold nanoparticles – cAuNPs; Polyvinylpyrrolidone
 499 coated gold nanoparticles – PVP-AuNPs; Au + GEM – 80 $\mu\text{g.L}^{-1}$ Au⁺ or AuNPs
 500 (cAuNPs or PVP-AuNPs) with 150 $\mu\text{g.L}^{-1}$ GEM.

501 **Table 1.** The relative percentage of effect in the different assessed endpoints, after 24 h liver organ culture single and combined
502 exposures to 80 µg.L⁻¹ gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs), 80 µg.L⁻¹ ionic gold
503 (Au⁺) and 150 µg.L⁻¹ gemfibrozil (GEM) compared with control. Observed (**O**) % in the combined exposures refers to measured effects and
504 the Predicted (**P**) % were derived by the sum of single exposure effects. *Significant differences to control (Dunnett’s test, *p*<0.05).

Assessed Endpoints	% of effect related to control										
	Au ⁺	7 nm cAuNPs	7 nm PVP-AuNPs	40 nm cAuNPs	40 nm PVP-AuNPs	GEM	Au ⁺ + GEM	7 nm cAuNPs + GEM	7 nm PVP-AuNPs + GEM	40 nm cAuNPs + GEM	40 nm PVP-AuNPs + GEM
Catalase Activity	40	27	- 131 *	-11	-45	6	P: 46 O: - 244 *	P: 33 O: - 213 *	P: - 124 O: - 136 *	P: - 5 O: - 132 *	P: - 39 O: - 204 *
Glutathione Reductase Activity	- 57	- 73	- 74	- 44	- 47	- 37	P: - 94 O: - 333 *	P: - 110 O: - 150	P: - 111 O: - 159	P: - 81 O: - 231	P: - 84 O: - 549 *
Glutathione S-Transferases Activity	- 69	- 54	- 80	- 71	- 53	- 4	P: - 73 O: - 7	P: - 58 O: - 53	P: - 84 O: - 22	P: - 74 O: - 8	P: - 57 O: -20
Lipid Peroxidation Levels	- 42	- 22	8	- 31	- 10	- 14	P: - 57 O: - 114	P: - 36 O: - 679	P: - 7 O: - 1048 *	P: - 45 O: - 896	P: - 24 O: - 2988 *
DNA Damage Index	- 128 *	- 104 *	- 115 *	- 26	- 112 *	- 162 *	P: - 290 O: - 145 *	P: -266 O: - 197 *	P: - 277 O: - 177 *	P: - 188 O: - 221 *	P: - 274 O: - 193 *

505 **4. Discussion**

506 In the present study, the agglomeration/aggregation of AuNPs found within 12
507 h in DMEM+FBS may influence the nanoparticles (NPs) toxicity. Several research
508 groups already reported that aggregation of NPs in cell culture media or PBS
509 might be prevented by adding serum, presumably due to proteins adsorbing onto
510 the particle surface (Allouni et al. 2009; Barreto et al. 2015; Balog et al. 2015;
511 Bihari et al. 2008; Mahl et al. 2010, 2012). Barreto et al. (2015) reported that
512 cAuNPs immediately aggregated/agglomerated in DMEM whereas in
513 DMEM+FBS they were stable for 12 hours. In the present study, at 0 h, in
514 DMEM+FBS, the size of AuNPs increased (although non-significantly), the SPR
515 peaks shifted toward higher wavelengths and ZP values were slightly less
516 negative, as previously reported (Barreto et al. 2015). This suggests that FBS
517 was bound to AuNPs, a relevant feature to take into consideration because, as
518 previously reported, the attachment of FBS with NPs may influence its
519 incorporation into the cells/organs and consequently reduce NPs toxic effects
520 (Durán et al. 2015). Considering the effect of AuNPs concentration on the
521 behaviour of the particles, it was observed that the time needed for AuNPs to
522 aggregate/agglomerate in the DMEM+FBS decreased with the increase of
523 AuNPs concentration. This is an expected finding because with the increasing
524 number of particles per volume, the probability of NPs collisions and consequent
525 agglomeration/aggregation will increase (Barreto et al. 2015). At 12 h, the
526 medium (DMEM+FBS with AuNPs) was bluer at the highest concentrations of
527 AuNPs, corroborating the previously described information. According to Zeng et
528 al. (2012), the surface energy of AuNPs increases with the decrease of the
529 diameter. Therefore, smaller AuNPs interact more strongly with the compounds

530 present in the solution, leading to size-dependent aggregation of AuNPs (Zeng
531 et al. 2012). In the present study, this was not visually observed. Additionally, at
532 12 h, all 7 and 40 nm AuNPs already had aggregated/agglomerated. Thus, for
533 the same concentration, the tested AuNPs sizes displayed similar behaviour in
534 the test media, in terms of aggregation/agglomeration and stability period.
535 However, different sizes of aggregates/agglomerates were detected, depending
536 on the initial size of AuNPs, with aggregates/agglomerates resultant from 40 nm
537 AuNPs being bigger than those resultant from 7 nm AuNPs.

538 Regarding Au concentration in the medium, at 24 h, a marked decrease was
539 observed after exposure to AuNPs. This may be explained by the
540 aggregation/agglomeration of the NPs and subsequent sedimentation of the
541 aggregates/agglomerates.

542 In the available literature, AuNPs toxicity data are often conflicting due to the
543 variability of the toxicity assays used in terms of cell lines, exposure times,
544 assessed endpoints, NPs concentrations and chemical/physical properties.
545 AuNPs have been reported as “nontoxic” according to some *in vitro* tests (Alkilany
546 and Murphy 2010; Connor et al. 2005; Luis et al. 2016; Shukla et al. 2005). Shukla
547 et al. (2005), using RAW264.7 macrophage murine cell line, reported that AuNPs
548 (size range from 3 to 8 nm; concentrations between 10 and 100 μM) are not
549 cytotoxic, reducing the production of reactive oxygen and nitrite species and not
550 eliciting secretion of proinflammatory cytokines, making them suitable candidates
551 for nanomedicine. Connor et al. (2005) reported that 18 nm AuNPs exposure did
552 not cause acute cytotoxicity in human K562 cells, at concentrations up to 250 μM .
553 Luis et al. (2016) also demonstrated *in vitro* that 7 nm AuNPs (concentrations
554 between 54 ng.L^{-1} to 2.5 mg.L^{-1}) did not affect *Mytilus galloprovincialis*

555 haemolymph' acetylcholinesterase nor gills' GST activities. However, other authors
556 have reported that AuNPs may present toxicity (Baharara et al. 2016; Goodman et
557 al. 2004; Li et al. 2010; Pan et al. 2009; Tkachenko et al. 2004). AuNPs (20 nm; 1
558 nM) induced autophagy with oxidative stress in MRC-5 human lung fibroblasts (Li
559 et al. 2010). The investigation of Baharara et al. (2016) demonstrated the induction
560 of apoptosis in human HeLa cell line treated with 100 and 400 $\mu\text{g}\cdot\text{mL}^{-1}$ AuNPs
561 (size range from 10 to 42 nm). HeLa and 3T3/NIH mouse embryo fibroblast cell
562 lines exposed to AuNPs (20 nm; 0.98 nM) presented decreased cell viabilities
563 (Tkachenko et al. 2004).

564 As previously reported, the possible adverse effects of AuNPs may be
565 attributed to: 1) their interaction with the cell membrane (Goodman et al. 2004);
566 2) oxidative stress leading to cytotoxicity effects (Pan et al. 2009); 3) the inhibition
567 of metabolic activity (e.g., leading to mitochondrial damage; Panessa-Warren et
568 al. 2008); 4) possible damage or alteration in the nuclear DNA (Panessa-Warren
569 et al. 2008; Schulz et al. 2012). In the current study, AuNPs induced oxidative
570 stress and damage to different cellular components (DNA strand breaks and lipid
571 membrane peroxidation), even at low concentrations ($4 \mu\text{g}\cdot\text{L}^{-1}$), with effects
572 dependent on the AuNPs size, coating and concentration. The 7 nm AuNPs
573 induced more effects than 40 nm AuNPs. Only 7 nm AuNPs increased LPO
574 levels. At $4 \mu\text{g}\cdot\text{L}^{-1}$, only 7 nm AuNPs caused DNA damage. This may be explained
575 by the higher ability of 7 nm AuNPs to be incorporated by the cell comparing with
576 40 nm AuNPs. An *in vivo* genotoxic effect of different sizes of AuNPs (2, 20 and
577 200 nm) was observed by Schulz et al. (2012) in the lungs of rats, with DNA
578 damage presenting a weak size-related increase of the mean tail intensity (Schulz
579 et al. 2012). As previously described, 15 nm AuNPs *in vitro* permeation on rat

580 skin was higher and more rapid than 102 and 198 nm AuNPs (Sonavane et al.
581 2008). In the present study, the 7 nm PVP-AuNPs, which presented the smallest
582 sizes during the experimental test, were those inducing more pronounced effects
583 in the liver. For instance, 7 nm PVP-AuNPs induced significantly higher levels of
584 peroxidative damage than cAuNPs. Comparing 40 nm cAuNPs and PVP-AuNPs,
585 the latter also induced effects (for instance DNA integrity loss) at concentrations
586 lower than those induced by cAuNPs. Previous studies also reported different
587 effects of AuNPs with different coatings (Iswarya et al. 2016; Fraga et al. 2013;
588 Paino et al. 2012). In a mice model, 96 h exposure to 65 nm PVP-AuNPs induced
589 more effects in the DNA of liver cells (assessed as DNA strand breaks) than 29
590 nm cAuNPs (Iswarya et al. 2016). In the present study, in general, the toxicity of
591 AuNPs was dependent on concentration of nanoparticles, with the highest tested
592 concentrations inducing more effects than the lowest tested concentrations,
593 regardless of the coating and size of AuNPs.

594 Comparing the present *in vitro* results with those obtained in *in vivo* exposures
595 of *S. aurata* to AuNPs (Barreto et al. 2019b, 2020), some similar trends were
596 observed, namely in terms of increases of CAT and GR activities, despite
597 differences in the exposure length (*in vitro*: 24 h and *in vivo*: 96 h). However,
598 some dissimilar results were also detected. For instance, GST activity was not
599 altered after *in vitro* exposures whereas *in vivo* exposure of 1,600 $\mu\text{g.L}^{-1}$ 40 nm
600 PVP-AuNPs increased hepatic GST activity (Barreto et al. 2019b). LPO levels
601 were not altered *in vitro* in the range of concentrations 4 to 1,600 $\mu\text{g.L}^{-1}$ although
602 1,600 $\mu\text{g.L}^{-1}$ of 7 nm cAuNPs increased *in vivo* liver LPO levels (Barreto et al.
603 2020). Another interesting finding is that 7 nm cAuNPs were the ones inducing
604 more adverse effects to *S. aurata* after *in vivo* exposure (Barreto et al. 2020)

605 whereas 7 nm PVP-AuNPs were the ones inducing more adverse effects to liver
606 organ culture of *S. aurata*. These dissimilar results may be due to
607 aggregation/agglomeration state of AuNPs dependent on the medium where they
608 are present and the time of exposure. In addition, in the *in vivo* tests the whole
609 living organism is used, and a range of mechanisms can occur in different
610 tissues/organs to protect/eliminate a contaminant whereas in the *in vitro* test, only
611 the mechanisms involved in cell, tissue or organ used are evaluated (i.e. the *in*
612 *vitro* models do not represent all the complexity found in an *in vivo* model).
613 Therefore, the results may not be always equivalent.

614 In the present work, gold in the nano form induced more adverse effects in the
615 liver organ culture of *S. aurata* than the ionic form. Oxidative damage was only
616 detected after the exposure to AuNPs. In agreement with the described results,
617 the study of Barbasz and Oćwieja (2016), using two types of human cell lines,
618 reported a higher cytotoxicity of AuNPs than Au⁺ (Barbasz and Oćwieja 2016).
619 However, other studies reported a higher toxicity of Au⁺ comparing with AuNPs
620 (Farkas et al. 2010; Luis et al. 2016). Farkas et al. (2010) reported that, in rainbow
621 trout (*Oncorhynchus mykiss*) hepatocyte cells, the *in vitro* exposure to 17.4 mg.L⁻¹
622 Au⁺ significantly increased ROS levels. At the same concentration, AuNPs did
623 not have any effect (Farkas et al. 2010). Luis et al. (2016) in another *in vitro* test
624 also showed that Au⁺ significantly decreased the haemolymph'
625 acetylcholinesterase and gills' GST activities of mussel *Mytilus galloprovincialis*.
626 However, no significant alterations were found after *in vitro* exposure to AuNPs,
627 regardless of their coating (Luis et al. 2016). A previous 96 h *in vivo* study with *S.*
628 *aurata* also showed the highest toxicity of Au⁺ comparing with AuNPs (Barreto et
629 al. 2020). There are few available studies about the possible mechanisms of Au⁺

630 toxic action. Nonetheless, the Au⁺ ability to undergo redox reactions with peptides
631 and proteins, particularly involving sulfur amino acids, to deprotonate and bind to
632 peptide amide bonds and cross-link histidine imidazole rings has been reported
633 (Best and Sadler 1996; Luis et al. 2016). Dissolution can play a critical role in the
634 fate, behaviour and toxicity of NPs. Some NPs can dissolve quickly in aqueous
635 media and the toxicity of some metal-based NPs, such as zinc oxide NPs and
636 silver NPs, results from the metal ions released. Contrary, AuNPs are insoluble
637 and its toxicity is not associated with the release of ions (Peng et al. 2017).

638 GEM exposure increased CAT and GR activities and LPO levels at the highest
639 tested concentration but affected the DNA integrity at all the tested
640 concentrations. *In vitro* toxicity of GEM was previously reported in the hepatoma
641 fish cell line PLHC-1 (Zurita et al. 2007) manifested through a reduced protein
642 content, neutral red uptake, methylthiazol metabolization and lysosomal function.
643 *In vivo*, 96 h of GEM exposure increased *S. aurata* hepatic CAT (at 15,000 µg.L⁻¹
644 ¹) and GR (from 15 to 15,000 µg.L⁻¹) activities and LPO levels (at 1.5 µg.L⁻¹)
645 (Barreto et al. 2018).

646 The effects of the concomitant exposure to AuNPs and GEM were, for many
647 endpoints, higher than the predicted. A previous 96 h *in vivo* study also showed
648 that the effects on *S. aurata* hepatic CAT and GR activities of the combined
649 exposures – 40 nm AuNPs with GEM – were higher than the sum of the effects
650 of each contaminant alone (Barreto et al. 2019b). The prediction of potential
651 synergistic effects between AuNPs and GEM, found in the present study, is a
652 relevant finding considering that, in the environment, there is a variety of
653 contaminants that may interact with each other. To our best knowledge, a single
654 *in vitro* study has so far assessed the combined effects of AuNPs and

655 pharmaceuticals (carbamazepine and fluoxetine) in aquatic organisms (Luis et al.
656 2016). It was demonstrated that AuNPs, in combined exposures, may
657 significantly alter the effects of the pharmaceuticals carbamazepine and
658 fluoxetine, even at concentrations that may be considered environmentally
659 relevant. These effects were dependent on the coating of NPs and tested
660 endpoint. In the present study, the detected effects of the combined exposures
661 were also dependent on the characteristics of AuNPs, with 40 nm PVP-AuNPs
662 with GEM inducing more synergistic effects than 40 nm cAuNPs combined with
663 GEM and 7 nm AuNPs plus GEM.

664 The liver organ culture of *Sparus aurata* was sensitive to low concentrations of
665 the tested contaminants and allowed to differentiate responses to NPs with
666 different characteristics: size and coating. They also allowed the study of
667 combined exposures, proving sensitive in discriminating experimental conditions.
668 Taking into account that the organ cultures involve “the maintenance or growth
669 of tissues, organ primordia or the whole or parts of an organ *in vitro* for a period
670 of 24 h or longer, in a way which may allow differentiation and/or preservation of
671 architecture and/or function” (Oliveira et al. 2003), this approach showed to be
672 very useful, supporting its use as an *in vitro* model. Further studies, analysing
673 different types of contaminants, are encouraged to understand if the liver organ
674 culture can be used as an alternative to *in vivo* testing.

675

676 **5. Conclusions**

677 The *in vitro* system used in the present study proved to be a valuable approach
678 to evaluate the single and combined effects of contaminants, such as
679 nanoparticles and pharmaceuticals, to aquatic organisms. Gold nanoparticles

680 (AuNPs) induced oxidative stress, increasing the activities of catalase and
681 glutathione reductase, and damage in DNA and cellular membranes, even at low
682 concentrations ($4 \mu\text{g.L}^{-1}$), in the liver organ culture of *Sparus aurata*. The effects
683 were dependent on the size, coating and concentration of AuNPs, being the 7 nm
684 PVP-AuNPs that induced higher effects. Gold in the nano form caused more
685 adverse effects than the ionic form of the metal. Additionally, gemfibrozil (GEM)
686 also induced DNA damage at an environmental relevant concentration (1.5
687 $\mu\text{g.L}^{-1}$). In many endpoints, the combined exposures of AuNPs and GEM induced
688 higher effects than the predicted, being an important finding considering that, in
689 the environment, there is a diversity of contaminants that may interact with each
690 other.

691

692 **Conflict of interest statement**

693 The authors declare that there are no conflicts of interest.

694

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709 **6. References**

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