# Biological effects and bioaccumulation of gold in gilthead seabream (Sparus aurata) – Nano versus ionic form

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2	Highlights
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4	<ul> <li>Gold (nano or ionic form) induced lipid peroxidation and genotoxicity on fish;</li> </ul>
5	<ul> <li>Ionic gold induced more adverse effects than a nano form of the metal;</li> </ul>
6	<ul> <li>Citrate coated nanoparticles were more bioactive than PVP coated</li> </ul>
7	nanoparticles.
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10	

#### 11 Abstract

The question of whether gold (Au) is more toxic as nanoparticles or in its 12 ionic form remains unclear and controversial. The present work aimed to clarify 13 the effects of 96 h exposure to 4, 80 and 1600  $\mu$ g.L<sup>-1</sup> of 7 nm gold nanoparticles 14 (AuNPs) - (citrate coated (cAuNPs) or polyvinylpyrrolidone coated (PVP-15 AuNPs)) - and ionic Au (iAu) on gilthead seabream (Sparus aurata). Effects at 16 17 different levels of biological organization (behaviour, neurotransmission, biotransformation, oxidative stress/damage and genotoxicity) were assessed. 18 cAuNPs induced oxidative stress and damage (lipid peroxidation increase), 19 20 even at 4  $\mu$ g.L<sup>-1</sup>, and reduced the ability of *S. aurata* to swim against a water flow at 1600 µg.L<sup>-1</sup>. Exposure to cAuNPs induced more adverse effects than 21 exposure to PVP-AuNPs. All tested concentrations of Au (nano or ionic form) 22 induced DNA breaks and cytogenetic damage in erythrocytes of S. aurata. 23 Generally, iAu induced significantly more effects in fish than the nano form, 24 25 probably associated with the significantly higher accumulation in the fish tissues. No fish mortality was observed following exposure to AuNPs, but 26 mortality was observed in the group exposed to 1600 µg.L<sup>-1</sup> of iAu. 27

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29 **Keywords:** nanotoxicity; gold; marine fish; seawater; biomarkers

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### 31 **1. Introduction**

Throughout its history, gold (Au) has been recognized as an inert, non-toxic and biocompatible noble metal with therapeutic properties (Daniel and Astruc 2004; Fratoddi et al. 2015). However, when Au decreases to nanometer

dimensions, the safety of the resulting nanomaterials has been questioned 35 36 (Boverhof et al. 2015). Gold nanoparticles (AuNPs) have been widely used in medicine and biological research (Fratoddi et al. 2015), including targeted 37 delivery of drugs (Ghosh et al. 2008), imaging and diagnosis (Bhattacharya and 38 Mukherjee 2008). Its application in aquaculture as antimicrobial agent (Saleh et 39 al. 2016) and to detect contaminants (Loganathan and John 2017) has also 40 been investigated. Despite the widespread use of AuNPs and consequent 41 release to the environment, there is a limited understanding of their 42 consequences for environmental health (Barreto et al. 2019a; Barreto et al. 43 44 2019b; Iswarya et al. 2016; Na-Phatthalung et al. 2018; Teles et al. 2016). In addition, the question of whether AuNPs are more toxic than ionic Au (iAu) 45 remains unresolved (Barbasz and Oćwieja 2016; Botha, James, and Wepener 46 47 2015; Dedeh et al. 2015; Farkas et al. 2010; Luis et al. 2016). Table 1 presents the example of several in vitro and in vivo studies where different toxicity 48 outputs were achieved. 49

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Table 1. Examples of studies assessing the toxicity of nano versus ionic
gold. Ref. – Reference; PVP – Polyvinylpyrrolidone; BSA – Bovine serum
albumin; ROS – Reactive oxygen species. 1 – Barbasz et al. (2016); 2 – Luis et
al. (2016); 3 – Botha et al. (2015); 4 – Dedeh et al. (2015); 5 – Farkas et al.
(2010).

Test type	Cell/ Organism	Exposure time	Endpoint/ Parameter	Coating	Size/Shape (nm)	Dose	More toxic: Ionic or nano form	Ref.
In vitro	Human promyelocytic cells of the HL-60 line Human histiocytic lymphoma cell line U- 937	24, 48 and 72 h	Cytotoxicity Nitric oxide and reduced glutathione levels	Tannic acid	Spherical 21	0.75 to 25 ppm	Nano	1
In	Mytilus	10 min	Enzymatic activities	Citrate, PVP and	Spherical	54 ng $\cdot L^{-1}$ to	Ionic	2

vitro	galloprovincialis hemolymph and subcellular fraction of gills			BSA	7	2.5 mg·L <sup>-1</sup>		
In vivo	Daphnia pulex, D. magna, Danio rerio, Poecilia reticulata, Labeobarbus aeneus, Pseudocrenilabrus philander, Tilapia sparrmanii, Oreochromis mossambicus	48 and 96 h	Species sensitivity distributions	Citrate	Spherical 14	0.0005 to 200 mg.L <sup>-1</sup>	lonic	3
In vivo	Danio rerio	20 d	Gene expression	Citrate	Spherical 14	0.25 and 0.8 µg.L <sup>-1</sup>	Nano	4
In vitro	Hepatocyte cell culture of Oncorhynchus mykiss	2 and 48 h	Cytotoxicity and ROS formation	Citrate	Spherical 5-10	0.063 to 19 mg.L <sup>−1</sup>	Ionic	5

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Thus, the aim of the present study was to add value to the shortage of 57 studies available comparing both Au forms and to investigate the effects of Au 58 on the top predator Sparus aurata after 96 h exposure to 7 nm AuNPs (citrate 59 coated (cAuNPs) or polyvinylpyrrolidone coated (PVP-AuNPs)) and iAu. To the 60 best of our knowledge, the present study is the first one testing the effects of 61 both forms of Au to a marine fish. AuNPs of small size were chosen due to the 62 63 reported highest effects attributed to small sizes (Coradeghini et al. 2013; Iswarya et al. 2016; Xia et al. 2017). Two coatings of AuNPs were tested to 64 65 clarify whether they determine the effects of AuNPs in the fish. Swimming performance, the activity of enzymes involved in neurotransmission 66 (cholinesterases - ChE), in biotransformation (glutathione S-transferases -67 GST) and antioxidant defence (glutathione reductase (GR), catalase (CAT) and 68 glutathione peroxidase (GPx)), non-enzymatic defence (non-protein thiols -69 NPT), oxidative damage (in DNA and cellular membranes), DNA strand breaks 70 and nuclear abnormalities were assessed. The concentration of Au was also 71

quantified in relevant tissues (gills, liver, spleen and muscle). The main specific
aims were: 1) to clarify which Au form is more toxic and bioaccumulative to this
marine fish (nano versus ionic); and 2) to clarify the effect of coating in the
AuNPs toxicity (cAuNPs versus PVP-AuNPs).

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# 77 2. Material and Methods

# 78 **2.1. Gold nanoparticles (AuNPs) – Synthesis and characterization**

cAuNPs with 7 nm diameter were synthesized based on the method 79 described by Shiba et al. (2013). The citrate reduction method, one of the most 80 widely used in AuNPs synthesis, was chosen due to the reasons described on 81 82 previous publications (Barreto et al. 2019a; Barreto et al. 2019b). PVP-AuNPs were obtained by coating part of cAuNPs with polyvinylpyrrolidone (PVP) as 83 described in detail by Barreto et al. (2015). PVP is a water-soluble, nontoxic and 84 biodegradable homopolymer. It is an excellent coating agent, especially for 85 86 noble metals NPs (Das et al. 2017; Min et al. 2009). This polymer is frequently used as AuNPs coating agent to increase its stability and to promote biological 87 interactions (Min et al. 2009). The characterisation of AuNPs stock suspensions 88 and AuNPs in the experimental media (artificial seawater - ASW) and in 89 ultrapure water was performed as described in previous publications (Barreto et 90 91 al. 2019a; Barreto et al. 2019b)

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#### 93 **2.2. Bioassay**

94 **2.2.1. Fish** 

Juvenile gilthead seabream (*Sparus aurata*) with length 7.6  $\pm$  0.1 cm, acquired from an aquaculture facility in Spain (Santander), were acclimated for

97 1 month in aquaria containing aerated and filtered artificial seawater (ASW, 98 prepared by dissolving the salt in reverse osmosis water to obtain a salinity of 99 30), under controlled temperature (17°C) and natural photoperiod. During this 100 period, the fish were fed daily at a ratio of 1 g per 100 g of fish with commercial 101 fish food (Sorgal, Portugal).

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# 103 2.2.2. Experimental design

During the bioassay, temperature, salinity, conductivity, pH, dissolved oxygen 104 and aeration conditions were similar to conditions of the acclimation period. The 105 106 experiment followed, in general, the OECD guideline (number 203) for fish acute bioassays (OECD 1992). Fish (n=12 per condition) were randomly 107 108 distributed in the experimental aquaria (3 per condition) in the ratio 1 g of fish 109 per 1 L of ASW and exposed for 96 h to the following experimental conditions: 0, 4, 80 and 1600 µg.L<sup>-1</sup> AuNPs (citrate and PVP coating) and iAu. The lowest 110 concentration tested (4  $\mu$ g.L<sup>-1</sup>) was a compromise between the predicted values 111 112 of AuNPs for the aquatic environment (0.14  $\mu$ g L<sup>-1</sup>) (García-Negrete et al. 2013; Tiede et al. 2009) and the potentially detectable Au concentration in fish tissues. 113 114 The other concentrations tested were 20-fold increases.

Part of the experimental media (approx. 80%) was renewed daily to prevent significant AuNPs alteration and to reduce the build-up of metabolic residues, after checking fish mortality and behaviour alterations and assessing the water parameters (temperature, salinity, pH and dissolved oxygen). Water samples were collected daily (at 0 and 24 h) from each experimental aquarium for the gold quantification. Water samples collected at 0 h, correspond to the water collected at the beginning of the assay and every time renewal of the media

took place (immediately after renewal). Water samples collected at 24 h,
correspond to the water collected 24 h after the beginning of the test and 24 h
after the renewal of the media.

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# 126 **2.3. Assessment of swimming performance**

After 96 h exposure, fish were individually introduced into a long flume and induced to swim against a water flow of 19 L.min<sup>-1</sup>. The time (in seconds) that fish spent swimming against the water flow was recorded. More information about this behavioural assessment can be found in a previous study of Barreto et al. (2019b).

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# 133 **2.4. Collection of biological material**

134 After a 2 h recovery period, fish were anesthetized with tricaine methanesulfonate (MS-222), blood samples were collected from the posterior 135 cardinal vein and then the animals were euthanized by spinal section. For the 136 comet assay, blood samples were diluted with saline phosphate buffer. Blood 137 smears were prepared for the assessment of erythrocytic nuclear abnormalities 138 (ENAs). Liver, gills, muscle and brain were removed from seven fish and stored 139 at -80°C until biochemical biomarkers analysis. Liver, gills, spleen and muscle 140 were taken from five animals and kept at -20°C until gold quantification. 141

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# 143 **2.4.1. Biochemical biomarkers analysis**

Liver and gills were homogenized in potassium phosphate buffer (0.1 mM, pH 7.4) using an ultrasonic homogenizer. The homogenate was then divided into three aliquots for: lipid peroxidation (LPO) assay, NPT quantification and

post-mitochondrial supernatant (PMS) preparation. To prevent oxidation, the aliquot of homogenate for LPO evaluation was transferred to a microtube with 4% BHT (2,6-di-tert-butyl-4-methylphenol) in methanol. The aliquots for LPO and NPT levels determination were stored at -80°C until analysis. PMS was accomplished by centrifugation (12 000 g for 20 min at 4°C) and aliquots were stored at -80°C until determination of GST, CAT, GPx and GR activities.

Muscle and brain tissues were homogenized in potassium phosphate buffer 153 (0.1 mM, pH 7.2). Part of the homogenate was transferred to a microtube with 154 4% BHT and stored at -80°C until LPO quantification. The remaining part was 155 centrifuged (3300 g for 3 min at 4°C), and the obtained supernatant was 156 collected and stored at -80°C until ChE activity determination. Protein 157 concentration of all the samples was determined according to Bradford (1976), 158 adapted to microplate, using bovine  $\gamma$ -globuline as standard. ChE activity was 159 determined according to the Ellman's method (1961) adapted to microplate 160 (Guilhermino et al. 1996). CAT activity was assayed as described by Claiborne 161 (1985). GR activity was estimated according the method of Carlberg and 162 163 Mannervik (1975) adapted to microplate (Lima et al. 2007). GPx activity was measured according to the method described by Mohandas et al. (1984), 164 modified by Athar and Iqbal (1998). NPT levels were determined based on the 165 method of Sedlak and Lindsay (1968), adopted by Parvez et al. (2003). GST 166 activity was determined by the method of Habig et al. (1974) adapted to 167 168 microplate (Frasco and Guilhermino 2002). LPO levels were assessed by the formation of thiobarbituric acid reactive substances (TBARS) based on Ohkawa 169 et al. (1979), adapted by Filho et al. (2001). More details on the biochemical 170 biomarkers analysis can be found on the study of Barreto et al. (2019b). 171

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# 173 **2.4.2.** Comet and erythrocytic nuclear abnormalities (ENAs) assays

The alkaline comet assay was conducted according to the method of Singh 174 et al. (1988) with some adaptations, as previously described by Barreto et al. 175 (2017). The sensitivity and specificity of the assay was improved by the 176 incubation of the lysed cells (nucleoids) with a lesion-specific endonuclease, 177 formamidopyrimidine DNA glycoslyase (Fpg). Fpg was chosen because it is a 178 protein recommended for the detection of oxidative DNA base damage, in 179 particular 8-OH guanine, as well as other damaged purines and abasic sites 180 181 (AP sites) and ring-opened N-7 guanine adducts (Albertini et al. 2000; Epe et al. 1993; Li, Laval, and B. Ludlum 1997; Speit et al. 2004; Tchou et al. 1994; Tice 182 et al. 2000; Tudek et al. 1998). The method for enzyme Fpg conjugated with 183 184 comet assay was performed according to procedures previously reported (Collins 2014; Collins et al. 1997). Two replicate comet slides were made for 185 each blood sample; one slide was treated with Fpg and the other without Fpg. A 186 positive control (fish blood treated with 25  $\mu$ M hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) for 10 187 min), with and without Fpg treatment, was also included in the assay. H<sub>2</sub>O<sub>2</sub> is a 188 189 recognized genotoxic agent, producing both strand breaks and oxidative DNA damage (Barreto et al. 2017; Termini 2000). After the lysis step, for the enzyme 190 treatment, the correspondent slides were removed from lysis buffer and were 191 washed 3 times in cold (4°C) enzymatic buffer solution (40 mM HEPES; 0.1 M 192 KCl; 0.5 mM EDTA; 0.2 mg.mL<sup>-1</sup> bovine serum albumin, pH 8.0). Fpg (45 μL, 193 1:60 diluted in enzymatic buffer solution) was added to the slides, which were 194 individually sealed with a coverslip and incubated during 30 min at 37°C. The 195 other steps involved in the comet assay were common to slides with or without 196

Fpg. Cells were classified according to tail length, into five classes (Collins 2004): class 0 – undamaged, without a tail; class 1 – with a tail shorter than the diameter of the nucleus; class 2 – with a tail length 1–2 times the diameter of the nucleus; class 3 – with a tail longer than twice the diameter of the nucleus; class 4 – comets with no nucleus. A damage index (DI), in arbitrary units, was assigned to each slide (for 100 cells) and consequently for each treatment, using the formula:

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$$DI = (0 \times n0) + (1 \times n1) + (2 \times n2) + (3 \times n3) + (4 \times n4)$$

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where: n = number of cells in each class. DI can range from 0 to 400 (de Andrade, de Freitas, and da Silva 2004).

The DNA damage index in cells treated with Fpg with the correspondent cells without the enzymatic treatment were compared to detect possible DNA oxidative damage.

The ENAs assay was carried out in mature peripheral erythrocytes according previous procedures and nuclear lesions were scored as micronuclei, lobed, segmented, kidney-shaped and vacuolated nuclei (Barreto et al. 2017; Pacheco and Santos 1996). Results were expressed as the ENAs frequency (‰) to each replicate (for 1000 cells) and consequently for each treatment using the formula:

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$$ENAs(\%) = \frac{\text{Number of cells containing ENAs}}{\text{Total number of cells counted}}$$

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# 220 2.5. Gold (Au) quantification

The determination of Au in the stock suspensions, in the experimental media 221 and fish tissues was performed according to the NIST NCL Method PCC-8 222 (NIST 2010). An iCAP<sup>™</sup> Q ICP-MS (inductively coupled plasma mass 223 spectrometry) instrument (Thermo Fisher Scientific, Bremen, Germany) was 224 used for the analysis. The ICP-MS instrumental conditions were as follow: 225 argon flow rate (14 L.min<sup>-1</sup>); auxiliary argon flow rate (0.8 L.min<sup>-1</sup>); nebulizer flow 226 rate (1.03 mL.min<sup>-1</sup>); RF power (1550 W) and dwell time (100 ms). The 227 elemental isotope <sup>197</sup>Au was monitored for analytical determination; <sup>159</sup>Tb and 228 <sup>209</sup>Bi were used as internal standards. The instrument was tuned daily for 229 230 maximum signal sensitivity and stability. More information about Au quantification can be found in the study of Barreto et al. (2019b). 231

232 Stock suspensions theoretical concentrations and number of nanoparticles 233 (NPs) were also estimated based on their UV-Vis spectra and sizes (Barreto et 234 al. 2015; Liu et al. 2007; Paramelle et al. 2014).

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# 236 2.6. Total gold (Au) content, bioaccumulation factor and estimated intake 237 for humans

Total Au content ([Au]<sub>total</sub>), in  $\mu$ g.g<sup>-1</sup>, was calculated, as described in a previous study from Barreto et al. (2019b), according to the formula:

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$$[Au]_{total} = [Au]_g + [Au]_l + [Au]_s + [Au]_{ms}$$

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Where  $[Au]_g$  is the concentration of Au in gills,  $[Au]_i$  the concentration of Au in liver,  $[Au]_s$  the concentration of Au in spleen and  $[Au]_{ms}$  the concentration of Au in muscle. The bioaccumulation factor (BAF), in L.g<sup>-1</sup>, was calculated according previous studies (Barreto et al 2019b; Yoo-Iam, Chaichana, and Satapanajaru 248 2014):

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# $BAF = [Au]_t / [Au]_{ASW}$

Where [Au]<sub>t</sub> is the content of Au in the specific fish tissue and [Au]<sub>ASW</sub> its concentration in the exposure media – ASW (collected daily at 0 h and quantified). More information about BAF calculation can be found in the study of Barreto et al. (2019b).

As *Sparus aurata* is a fish for human consumption an extrapolation of Au intake for humans was calculated, using the following formula (Barreto et al 2019b; Vieira et al. 2015; WHO 2008):

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# $\label{eq:autof} \textit{Au intake} = \frac{\textit{Amount of fish ingested * Au content in the ingested fish}}{\textit{Kilograms body weight}}$

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A human body weight of 60 kg was assumed (IPCS 2004) and the average 261 262 amount of fish ingested by each Portuguese person per year was set at 59 kg (Failler et al. 2007; Vieira et al. 2015). Au content in the ingested fish 263 corresponds to the content of Au determined in the fish muscle (µg.g<sup>-1</sup>). The 264 calculated Au intake values were compared with the maximum amount of Au 265 that each person may be exposed daily over their lifetimes without considerable 266 267 health risk – "tolerable daily intake" (TDI). This value was previously calculated and in detail explained in the study of Barreto et al. (2019b): 322 µg.kg<sup>-1</sup>. 268

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# 270 2.7. Statistical analysis

Data were tested for normality (Shapiro-Wilk test) and homogeneity of variance (Levene's test) using the Sigma Plot software package (SigmaPlot 12.0, 2011). Differences between treatments and control and between all the treatments were analysed using one-way analysis of variance (ANOVA), followed by Dunnett's or Tukey's test whenever applicable. Significant differences were assumed for p<0.05.

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278 **3. Results** 

# **3.1. Gold nanoparticles (AuNPs) – Characterisation and behaviour**

280 The synthetized cAuNPs presented a well-defined surface plasmon 281 resonance (SPR) peak – Figure 1A. Dynamic light scattering (DLS) analysis showed an average hydrodynamic size of the particles of 7 nm and a strongly 282 negative surface charge (-43 mV). Transmission electron microscopy (TEM) 283 analysis confirmed that almost all cAuNPs presented spherical shape and the 284 particles had similar sizes between them (Figure 1B). There was a slight shift in 285 286 the SPR peak to a longer wavelength for PVP-AuNPs (521 nm) when compared with cAuNPs (519 nm). DLS measurements showed a size of around 8 nm and 287 a less negative ZP than cAuNPs (-13 mV). Scanning electron microscopy 288 (SEM) analysis allowed the visualization of a PVP layer around some AuNPs 289 metal core (Figure 1C). 290



Figure 1. UV–Vis spectra (A), transmission electron microscopy image of citrate coated gold nanoparticles – cAuNPs (B) and scanning electron microscopy image of polyvinylpyrrolidone coated – PVP-AuNPs (C).

In the experimental media (ASW), 80 and 1600 µg.L<sup>-1</sup> cAuNPs changed the 295 296 colour from red to light blue, as a result of NPs agglomeration/aggregation, whereas PVP-AuNPs did not show colour alteration. At 4 µg.L<sup>-1</sup>, it was not 297 possible to detect any colour change. Moreover, the hydrodynamic size of 298 299 cAuNPs (1600 µg.L<sup>-1</sup>) in ASW at 0 h increased to around 160 nm, maintaining this size till the end of the test (96 h) - Table 2. The characteristic surface 300 plasmon resonance (SPR) peak detected in ultrapure water was not detected in 301 ASW (Table 2). Additionally, different peaks corresponding to different charges 302 were found in the zeta potential (ZP) analysis of the cAuNPs in ASW. Within 24 303 304 h, in the aquaria containing 1600  $\mu$ g.L<sup>-1</sup> of cAuNPs, a dark layer was visible as a 305 consequence of the sedimentation of the NPs aggregates/agglomerates. PVP-AuNPs (1600 µg.L<sup>-1</sup>) in ASW had similar characteristics (e. g. hydrodynamic 306 307 size, UV-Vis spectra and ZP) as the PVP-AuNPs in ultrapure water (Table 2). At 4 and 80 µg.L<sup>-1</sup>, it was not possible characterise the AuNPs because of the 308 detection limits of the techniques used. 309

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Table 2. Characteristics of gold nanoparticles (AuNPs), at 1600 µg.L<sup>-1</sup>, in
 ultrapure water and artificial seawater after 96 h. cAuNPs – Citrate coated gold
 nanoparticles; PVP-AuNPs – Polyvinylpyrrolidone coated gold nanoparticles;
 PdI – Polydispersity Index; SPR – Surface Plasmon Resonance; ZP – Zeta
 Potential; N. D. – Not detected

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	Size (nm)	Pdl	SPR (nm)	ZP (mV)				
	Ultrapure water							
cAuNPs	6.7	0.5	519.0	-43.3				
PVP-AuNPs	7.8	0.5	521.0	-12.8				
	Artificial seawater							
cAuNPs	159.8	0.8	N. D.	N. D.				
PVP-AuNPs	8.1	0.5	521.4	-12.6				

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# 319 **3.2. Gold (Au) quantification in experimental media**

AuNPs theoretical concentration versus measured concentrations (by ICP-MS) and the number of particles present in the AuNPs stock suspensions are shown in Table S1. The nominal versus measured concentrations of Au in the experimental media are presented in Table 3.

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Table 3. Nominal versus measured Au concentrations ( $\mu$ g.L<sup>-1</sup>) in experimental media (artificial seawater) at 0 and 24 h, after exposure to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results are expressed as mean ± standard error. <d.I. – Below the detection limit.

Nominal concentrations (µg.L <sup>-1</sup> )	Time (h)	Measuro cAuNPs	ed Au concentratior PVP-AuNPs	ns (µg.L <sup>⁻1</sup> ) Ionic gold
0	0	<d.l.< th=""><th><d.l.< th=""><th><d.l.< th=""></d.l.<></th></d.l.<></th></d.l.<>	<d.l.< th=""><th><d.l.< th=""></d.l.<></th></d.l.<>	<d.l.< th=""></d.l.<>
	24	<d.l.< td=""><td><d.l.< td=""><td><d.l.< td=""></d.l.<></td></d.l.<></td></d.l.<>	<d.l.< td=""><td><d.l.< td=""></d.l.<></td></d.l.<>	<d.l.< td=""></d.l.<>
4	0	2.4 ± 3.8	$7.5 \pm 0.7$	7.1 ± 0.4
	24	1.6 ± 2.8	$6.4 \pm 0.6$	7.1 ± 0.6
80	0	24.1 ± 1.1	50.0 ± 2.8	92.5 ± 0.9
	24	7.0 ± 0.6	38.4 ± 1.6	89.1 ± 1.3
1600	0	88.9 ± 7.0	1341.1 ± 51.7	1370.2 ± 36.0
	24	34.6 ± 5.8	1140.7 ± 19.9	1285.1 ± 81.8

At 0 h, the Au quantified in ASW, in general, was lower than the nominal 330 concentrations, with exception to 4 µg.L<sup>-1</sup> of PVP-AuNP and iAu (4 and 80 µg.L<sup>-1</sup> 331 <sup>1</sup>). The difference between the nominal and measured concentrations was more 332 noticeable in the case of the exposures to cAuNPs (Table 3). With the 333 increasing cAuNPs concentration, the difference between the nominal and 334 measured concentration of Au also increased. For the nominal concentration of 335 4 µg.L<sup>-1</sup> cAuNPs, the measured concentration of Au was 41% lower than the 336 expected. For PVP-AuNPs and iAu, the determined concentrations of Au were 337 88 and 78% higher than the expected, respectively. For the 80  $\mu$ g.L<sup>-1</sup> treatment, 338 339 the detected concentrations of Au in ASW were 70 and 38% lower than the nominal concentrations, after cAuNPs and PVP-AuNPs exposures, respectively. 340 For 80 µg.L<sup>-1</sup> of iAu, the measured concentration of Au was 16% higher than the 341 342 expected. At 1600 µg.L<sup>-1</sup>, the concentration of Au was 84, 16 and 14% lower than the expected for cAuNPs, PVP-AuNPs and iAu, respectively. The levels of 343 Au at 0 and 24 h decreased more for cAuNPs than PVP-AuNPs (Table 3). In 344 the nominal concentration 4 µg.L<sup>-1</sup>, an Au decrease of 33 and 15% was found 345 for cAuNPs and PVP-AuNPs, respectively. Concerning iAu, the measured 346 concentration at 0 h was similar to the measured at 24 h. In the nominal 347 concentration 80 µg.L<sup>-1</sup>, after 24 h of exposure, the concentrations of Au 348 decreased by 71, 23 and 4% for cAuNPs, PVP-AuNPs and iAu, respectively. 349 For the nominal concentration 1600  $\mu$ g.L<sup>-1</sup>, a decrease of Au concentration 350 after 24 h was also observed with 61% for cAuNPs, 15% for PVP-AuNPs and 351 6% for iAu. 352

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# **354 3.3. Biological responses of fish after exposure to gold (Au)**

After 24 h of exposure, one fish died in the 1600  $\mu$ g.L<sup>-1</sup> iAu treatment. As shown in Figure 2, the ability of *Sparus aurata* to continue swimming against a water flow was significantly decreased (p<0.05; Dunnett's test) when fish were exposed to 1600  $\mu$ g.L<sup>-1</sup> of cAuNPs and iAu.



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Figure 2. Resistance of *Sparus aurata* to withstand swimming against a water flow after 96 h exposure to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results are expressed as mean time in seconds  $\pm$  standard error. \*Significant differences to control (Dunnett's test, p<0.05).

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ChE activity in brain and muscle was not significantly altered by the exposure to the both forms of Au (p>0.05; ANOVA; Figure 3).



Figure 3. Brain (A) and muscle (B) cholinesterases (ChE) activity of *Sparus aurata* after 96 h exposure to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results are expressed as mean ± standard error.

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Concerning the enzymatic antioxidant defence, iAu (4 and 1600  $\mu$ g.L<sup>-1</sup>) significantly increased gills CAT activity (p<0.05; Dunnett's test; Figure 4A), whereas, liver CAT activity was not significantly affected by the exposure to both forms of Au (p>0.05; ANOVA; Figure 4B). PVP-AuNPs, at 80  $\mu$ g.L<sup>-1</sup>, significantly decreased the gills CAT activity (p<0.05; Dunnett's test; Figure 4A).





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GR activity in gills was significantly increased after exposure to 1600  $\mu$ g.L<sup>-1</sup> of cAuNPs (p<0.05; Dunnett's test; Figure 5A) whereas in the liver, GR activity was significantly increased by 80 and 1600  $\mu$ g.L<sup>-1</sup> of cAuNPs (p<0.05; Dunnett's test; Figure 5B). PVP-AuNPs, 80 and 1600  $\mu$ g.L<sup>-1</sup>, significantly decreased gills GR activity (p<0.05; Dunnett's test; Figure 5A).



Figure 5. Gills (A) and liver (B) glutathione reductase (GR) activity of Sparus 395 aurata after 96 h exposure to gold nanoparticles (citrate coated - cAuNPs and 396 polyvinylpyrrolidone coated - PVP-AuNPs) and ionic gold. Results are 397 expressed as mean ± standard error. \*Significant differences to control 398 (Dunnett's test, p<0.05). Different letters correspond to significant differences 399 between the treatments of each type of AuNPs and ionic form (Tukey's test, 400 401 p<0.05). Different numbers correspond to significant differences between each 402 treatment within the same concentration (Tukey's test, p<0.05).

404 cAuNPs and iAu at 4 and 80  $\mu$ g.L<sup>-1</sup>, respectively, significantly increased gills 405 GPx activity (p<0.05; Dunnett's test; Figure 6A). In the liver, 80  $\mu$ g.L<sup>-1</sup> of PVP-406 AuNPs decreased the GPx activity (p<0.05; Dunnett's test; Figure 6B).

403



Figure 6. Gills (A) and liver (B) glutathione peroxidase (GPx) activity of 408 Sparus aurata after 96 h exposure to gold nanoparticles (citrate coated -409 cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results 410 are expressed as mean ± standard error. \*Significant differences to control 411 (Dunnett's test, p<0.05). Different letters correspond to significant differences 412 between the treatments of each type of AuNPs and ionic form (Tukey's test, 413 p<0.05). Different numbers correspond to significant differences between each 414 treatment within the same concentration (Tukey's test, p<0.05). 415

416

Concerning the non-enzymatic antioxidant defence, all tested concentrations of cAuNPs and 1600  $\mu$ g.L<sup>-1</sup> of iAu significantly increased gills NPT levels (p<0.05; Dunnett's test; Figure 7A). In liver, only 1600  $\mu$ g.L<sup>-1</sup> of cAuNPs significantly increased the levels of NPT (p<0.05; Dunnett's test; Figure 7B).



422 Figure 7. Gills (A) and liver (B) non-protein thiols (NPT) levels of Sparus 423 aurata after 96 h exposure to gold nanoparticles (citrate coated – cAuNPs and 424 polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results are expressed as mean ± standard error. \*Significant differences to control 425 (Dunnett's test, p<0.05). Different letters correspond to significant differences 426 between the treatments of each type of AuNPs and ionic form (Tukey's test, 427 428 p<0.05). Different numbers correspond to significant differences between each treatment within the same concentration (Tukey's test, p<0.05). 429

430

431 cAuNPs (4  $\mu$ g.L<sup>-1</sup>) and iAu (1600  $\mu$ g.L<sup>-1</sup>) significantly increased gills GST 432 activity (p<0.05; Dunnett's test; Figure 8A). In liver, 1600  $\mu$ g.L<sup>-1</sup> iAu significantly 433 increased the activity of this enzyme (p>0.05; ANOVA; Figure 8B).

434



Figure 8. Gills (A) and liver (B) glutathione S-transferases (GST) activity of 436 Sparus aurata after 96 h exposure to gold nanoparticles (citrate coated -437 cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results 438 are expressed as mean ± standard error. \*Significant differences to control 439 (Dunnett's test, p<0.05). Different letters correspond to significant differences 440 between the treatments of each type of AuNPs and ionic form (Tukey's test, 441 p<0.05). Different numbers correspond to significant differences between each 442 treatment within the same concentration (Tukey's test, p<0.05). 443

444

As shown in Figure 9A, oxidative damage (assessed as TBARS levels) was found in gills after the exposure to all tested concentrations of cAuNPs (p<0.05; Dunnett's test). In liver, LPO levels significantly increased after the exposure to 1600 µg.L<sup>-1</sup> of cAuNPs (p<0.05; Dunnett's test; Figure 9B).



449

Figure 9. Gills (A) and liver (B) lipid peroxidation (LPO) levels of Sparus 450 aurata after 96 h exposure to gold nanoparticles (citrate coated - cAuNPs and 451 polyvinylpyrrolidone coated - PVP-AuNPs) and ionic gold. Results are 452 expressed as mean ± standard error. \*Significant differences to control 453 (Dunnett's test, p<0.05). Different letters correspond to significant differences 454 between the treatments of each type of AuNPs and ionic form (Tukey's test, 455 p<0.05). Different numbers correspond to significant differences between each 456 treatment within the same concentration (Tukey's test, p<0.05). 457

458

Increase of LPO levels was also detected in brain following exposure to 4 and 80  $\mu$ g.L<sup>-1</sup> of iAu (p<0.05; Dunnett's test; Figure 10A) and in muscle after the exposure to 80  $\mu$ g.L<sup>-1</sup> of iAu (p<0.05; Dunnett's test; Figure 10B).



Figure 10. Brain (A) and muscle (B) lipid peroxidation (LPO) levels of Sparus 464 aurata after 96 h exposure to gold nanoparticles (citrate coated - cAuNPs and 465 polyvinylpyrrolidone coated - PVP-AuNPs) and ionic gold. Results are 466 expressed as mean ± standard error. \*Significant differences to control 467 (Dunnett's test, p<0.05). Different letters correspond to significant differences 468 between the treatments of each type of AuNPs and ionic form (Tukey's test, 469 p<0.05). Different numbers correspond to significant differences between each 470 treatment within the same concentration (Tukey's test, p<0.05). 471

472

All the treatments induced genotoxic effects (p<0.05; Dunnett's test), 473 assessed by DNA strand breakage - Table 4. A dose response pattern was 474 found, with damage index increasing with the increase of Au concentration 475 (both forms). A DNA damage index around 319 was detected in animals 476 exposed to 1600 µg.L<sup>-1</sup> of iAu, the highest value detected considering all the 477 478 treatments (Table 4). In terms of damage classes, as shown in Table 4, the most abundant classes in the negative control group were class 0 and 1. Class 479 2 was the most detected in the exposures to 4  $\mu$ g.L<sup>-1</sup> and classes 2 and 3 in the 480 exposures to 80 µg.L<sup>-1</sup> (p<0.05; Dunnett's test). At 1600 µg.L<sup>-1</sup>, Au exposures 481 induced a DNA damage classified, mostly, in classes 3 and 4 (p<0.05; 482 483 Dunnett's test). No significant oxidative DNA damage was found (p>0.05; ANOVA). Comparing the DNA damage index in cells treated with Fpg with the 484 correspondent cells without the enzymatic treatment, no significant differences 485 were found (p>0.05; ANOVA) – Table 4. However, comparing the DNA damage 486 index in cells treated with H<sub>2</sub>O<sub>2</sub> with and without treatment with Fpg, in the cells 487 with Fpg the DNA damage index was significantly higher than those without Fpg 488 (p<0.05; Tukey's test). 489

490

Table 4. DNA damage classes, measured by the comet assay, of peripheral 491 blood cells from Sparus aurata after 96 h exposure to gold nanoparticles (citrate 492 coated - cAuNPs and polyvinylpyrrolidone coated - PVP-AuNPs) and ionic 493 gold. \*Significant differences to control (Dunnett's test, p<0.05); data are 494 presented as mean ± standard error. Different letters correspond to significant 495 differences between the treatments of each type of AuNPs and ionic form 496 (Tukey's test, p<0.05). A. U. – Arbitrary units; Fpg – Formamidopyrimidine DNA 497 glycoslyase. 498

4	g	g
-	-	-

Treatment		DNA da	amage clas	sses (%)		DNA	DNA damage	
group	0	1	2	2 3		index (A. U.)	with Fpg	
Control	35.9±2.6	58.2±2.9	5.7±0.9	0.2±0.1	-	70.1±2.5	69.9±4.2	
4 μg.L <sup>-1</sup> cAuNPs	0.8±0.4*	37.2±3.5*	47.6±4.8*	14.0±2.7*	0.4±0.2	176.0±4.0* <sup>A</sup>	192.0±2.6* <sup>A,1</sup>	
80 µg.L cAuNPs	0.2±0.2*	8.8±4.2*	42.0±2.6*	40.0±2.7*	9.0±0.5	248.8±6.9* <sup>B</sup>	261.8±2.9* <sup>B</sup>	
1600 µg.L <sup>-1</sup> cAuNPs	0.6±0.4*	11.4±4.4*	31.6±3.2*	32.4±2.5*	30.0±5.0*	291.8±15.3* <sup>B</sup>	301.4±13.8* <sup>B</sup>	
4 µg.L <sup>-1</sup> PVP-AuNPs	1.0±0.4*	28.8±3.9*	52.0±3.6*	16.2±2.0*	2.0±1.8	189.4±6.3* <sup>A</sup>	196.0±6.3* <sup>A,1,2</sup>	
80 µg.L <sup>-1</sup> PVP-AuNPs	0.2±0.2*	9.8±3.9*	31.0±2.5*	49.6±5.0*	7.8±2.1	251.8±7.7* <sup>B</sup>	257.4±7.2* <sup>B</sup>	
1600 µg.L <sup>-1</sup> PVP-AuNPs	0.6±0.4*	11.2±4.5*	25.6±1.7*	37.8±2.8*	30.8±4.9*	299.0±14.2* <sup>c</sup>	305.2±12.6* <sup>c</sup>	
4 µg.L <sup>⁻1</sup> ionic gold	0.8±0.4*	16.2±5.1*	50.2±2.9*	27.6±4.8*	5.2±2.5	220.2±13.7* <sup>A</sup>	240.2±7.9* <sup>A,2</sup>	
80 µg.L <sup>-1</sup> ionic gold	0.2±0.2*	11.4±4.4*	31.6±3.2*	32.4±2.5*	30.0±5.0*	291.8±15.3* <sup>B</sup>	290.6±13.0* <sup>B</sup>	
1600 µg.L <sup>-1</sup> ionic gold	-*	1.0±0.8*	18.4±3.1*	41.6±5.1*	39.0±2.1*	318.6±3.1* <sup>B</sup>	322.4±4.0* <sup>B</sup>	

500

All the treatments, with the exception to 4  $\mu$ g.L<sup>-1</sup> of cAuNPs and iAu, led to significantly higher ENAs frequency (p<0.05; Dunnett's test), as shown in Figure 11. The frequency of ENAs increased with the increase of Au concentration (both forms).



506

Figure 11. Erythrocytic nuclear abnormalities (ENAs) frequency in *Sparus* aurata after 96 h exposure to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. Results are expressed as mean  $\pm$  standard error. \*Significant differences to control (Dunnett's test, p<0.05). Different letters correspond to significant differences between the treatments of each type of AuNPs and ionic form (Tukey's test, p<0.05).

514

As shown in Table 5, the lobed nuclei abnormality was the most commonly detected in all the treatments, followed by kidney-shaped nuclei (p<0.05; Dunnett's test). The segmented, micronuclei and vacuolated nuclei abnormalities were the less detected – Table 5.

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Table 5. Erythrocytic nuclear abnormalities (ENAs) detected in *Sparus aurata* after 96 h exposure to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold. \*Significant differences to control (Dunnett's test, p<0.05); data are presented as mean  $\pm$ standard error. K – kidney-shaped nuclei; S – segmented nuclei; L – lobed nuclei; V – vacuolated nuclei; MN – micronuclei.

- 526
- 527

	ENAs frequency (‰)						
Treatment group	к	S	L	v	MN		
Control	0.8±0.2	0.1±0.1	1.5±0.5	0.1±0.1	0.0±0.0		
4 μg.L <sup>-1</sup> cAuNPs	2.0±0.8	0.3±0.2	2.3±0.4	0.3±0.2	0.0±0.0		
80 µg.L <sup>-1</sup> cAuNPs	4.0±1.1*	0.2±0.2	7.7±2.5*	0.0±0.0	-0.0±0.0		
1600 µg.L <sup>-1</sup> cAuNPs	7.6±1.3*	0.1±0.1	15.4±3.7*	0.0±0.0	0.0±0.0		
4 μg.L <sup>-1</sup> PVP-AuNPs	4.3±0.7*	0.0±0.0-	7.2±1.1*	0.0±0.0	0.0±0.0		
80 µg.L <sup>⁻1</sup> PVP-AuNPs	4.6±0.8*	0.4±0.2	12.2±3.8*	0.0±0.0	0.0±0.0		
1600 µg.L <sup>-1</sup> PVP- AuNPs	10.3±1.6*	0.0±0.0	18.3±4.1*	0.0±0.0	0.2±0.2		
4 μg.L <sup>-1</sup> ionic gold	3.3±0.5*	0.4±0.2*	6.3±1.4*	0.0±0.0	0.0±0.0		
80 µg.L <sup>-1</sup> ionic gold	4.7±8*	0.3±0.2	9.7±0.9*	0.3±0.2	0.0±0.0		
1600 μg.L <sup>-1</sup> ionic gold	4.5±1.0*	0.3±0.3	14.2±1.2*	0.5±0.3	0.0±0.0		

528

3.4. Total gold (Au) content, bioaccumulation factor and estimated intake
 for humans

Au did not accumulate significantly in the assessed tissues of *S. aurata* (p>0.05; ANOVA; Table 6) after the exposure to AuNPs. However, Au significantly accumulated in gills, liver and spleen of *S. aurata* after the exposure to 1600  $\mu$ g.L<sup>-1</sup> iAu (p<0.05; Dunnett's test; Table 6). The highest calculated BAF value (2 L.g<sup>-1</sup>) was found for the nominal concentration exposure of 4  $\mu$ g.L<sup>-1</sup> cAuNPs, in the spleen (Table 6).

**Table 6.** Gold content in tissues (gills, liver, spleen and muscle) of *Sparus aurata* exposed to gold nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-AuNPs) and ionic gold for 96 h and respective estimated bioaccumulation factor (BAF). Results are expressed as mean  $\pm$  standard error. \*Significant differences to control (Dunnett's test, p<0.05). <d.l. – Below the detection limit.

Nominal concentrations (µg.L <sup>-1</sup> )	Tissues	Go cAuNPs	ld content ( PVP-AuNP:	µg.g <sup>-1</sup> ) s lonic gold	cAuNPs	BAF (L.g <sup>-1</sup> ) PVP-AuNPs	lonic gold
0	Gills Liver Spleen Muscle	<d.l. <d.l. <d.l. <d.l.< th=""><th><d.l. <d.l. <d.l. <d.l.< th=""><th><d.l. <d.l. <d.l. <d.l.< th=""><th>- - -</th><th>- - -</th><th>- - -</th></d.l.<></d.l. </d.l. </d.l. </th></d.l.<></d.l. </d.l. </d.l. </th></d.l.<></d.l. </d.l. </d.l. 	<d.l. <d.l. <d.l. <d.l.< th=""><th><d.l. <d.l. <d.l. <d.l.< th=""><th>- - -</th><th>- - -</th><th>- - -</th></d.l.<></d.l. </d.l. </d.l. </th></d.l.<></d.l. </d.l. </d.l. 	<d.l. <d.l. <d.l. <d.l.< th=""><th>- - -</th><th>- - -</th><th>- - -</th></d.l.<></d.l. </d.l. </d.l. 	- - -	- - -	- - -
4	Gills Liver Spleen Muscle	<d.l. 0.1 ± 0.0 4.8 ± 0.4 <d.l.< th=""><th>0.2 ± 0.0 <d.l. <d.l. <d.l.< th=""><th>0.1 ± 0.0 0.1 ± 0.0 0.4 ± 0.0 <d.l.< th=""><th>0.0 2.0 -</th><th>0.0 - - -</th><th>0.0 0.0 0.1 -</th></d.l.<></th></d.l.<></d.l. </d.l. </th></d.l.<></d.l. 	0.2 ± 0.0 <d.l. <d.l. <d.l.< th=""><th>0.1 ± 0.0 0.1 ± 0.0 0.4 ± 0.0 <d.l.< th=""><th>0.0 2.0 -</th><th>0.0 - - -</th><th>0.0 0.0 0.1 -</th></d.l.<></th></d.l.<></d.l. </d.l. 	0.1 ± 0.0 0.1 ± 0.0 0.4 ± 0.0 <d.l.< th=""><th>0.0 2.0 -</th><th>0.0 - - -</th><th>0.0 0.0 0.1 -</th></d.l.<>	0.0 2.0 -	0.0 - - -	0.0 0.0 0.1 -
80	Gills Liver Spleen Muscle	$1.3 \pm 0.1$ $0.1 \pm 0.0$ $0.5 \pm 0.1$ $0.1 \pm 0.0$	0.2 ± 0.1 <d.l. 0.4 ± 0.0 0.1 ± 0.0</d.l. 	$1.2 \pm 0.4$ $0.8 \pm 0.4$ $1.6 \pm 1.0$ $0.2 \pm 0.0$	0.1 0.0 0.0 0.0	0.0 - 0.0 0.0	0.0 0.0 0.0 0.0
1600	Gills Liver Spleen Muscle	$3.3 \pm 0.3$ $0.5 \pm 0.0$ $3.3 \pm 1.9$ $0.1 \pm 0.0$	0.8 ± 0.2 2.4 ± 2.1 1.1 ± 1.0 <d.l.< th=""><th><math>8.2 \pm 4.5 *</math> <math>8.4 \pm 2.5 *</math> <math>6.4 \pm 4.0 *</math> <math>1.1 \pm 0.6</math></th><th>0.0 0.0 0.0 0.0</th><th>0.0 0.0 0.0 -</th><th>0.0 0.0 0.0 0.0</th></d.l.<>	$8.2 \pm 4.5 *$ $8.4 \pm 2.5 *$ $6.4 \pm 4.0 *$ $1.1 \pm 0.6$	0.0 0.0 0.0 0.0	0.0 0.0 0.0 -	0.0 0.0 0.0 0.0

543

544 The highest [Au]<sub>total</sub> value (around 24  $\mu$ g.g<sup>-1</sup>) was detected after exposure to

545 1600 μg.L<sup>-1</sup> iAu (p<0.05; Dunnett's test).

546

547 The highest estimated value for Au intake by a Portuguese citizen would be

for the condition 1600  $\mu$ g.L<sup>-1</sup> of iAu (Table 7).

549

**Table 7**. Estimated gold intake (µg per kg body weight per year), by each Portuguese person, after the ingestion of *Sparus aurata*, taking into account the total content of gold detected in muscle of fish after 96 h exposure to gold 553 nanoparticles (citrate coated – cAuNPs and polyvinylpyrrolidone coated – PVP-

554 AuNPs) and ionic gold.

555

Nominal concentrations (µg.L <sup>-1</sup> )	Estimated gold intake (µg.kg body weight per year) cAuNPs PVP-AuNPs Ionic gold					
4	-	-	-			
80	0.05	0.05	0.15			
1600	0.14	0.03	1.10			

# 556

### 557 4. Discussion

The coating of 7 nm AuNPs with PVP resulted in a slight shift in the SPR 558 559 peak to a longer wavelength when compared with the original cAuNPs as previously observed for the same AuNPs (Barreto et al. 2015). DLS 560 561 measurements showed an increased size of PVP-AuNPs and a less negative 562 ZP value when compared with cAuNPs, also in agreement with a previous study (Barreto et al. 2015). The detected size difference may be explained by the fact 563 that PVP presenting a larger size than citrate (Iswarya et al. 2016; Tejamaya et 564 565 al. 2012). In terms of ZP, the observed difference between cAuNPs and PVP-AuNPs may be explained by the fact that PVP is an uncharged molecule thus 566 making the PVP-AUNPs less negative than cAuNPs (Mahl et al. 2010). In the 567 experimental media (ASW), 80 and 1600 µg.L<sup>-1</sup> cAuNPs changed the colour, as 568 a result of NPs agglomeration/aggregation. PVP-AuNPs, at 80 and 1600 µg.L<sup>-1</sup>, 569 570 did not show colour alteration in ASW. These results are in agreement with the previous study of Barreto et al. (2015) which demonstrated that 7 nm PVP-571

AuNPs were stable in ASW for more than 30 days. Thus, the present study 572 573 confirmed that PVP-AuNPs may remain stable in suspension in a nano size range in ASW, whereas cAuNPs immediately alter their characteristics and 574 aggregate/agglomerate, increasing their size to more than 100 nm. These 575 characteristics and behaviour of different AuNPs may influence their 576 accumulation and effects to the organisms. NPs size may affect its 577 bioavailability to the organisms. When aggregates become too large for direct 578 transport across the cell membrane, uptake may be prevented (Vale et al. 579 2016). 580

Although the stability of the tested AuNPs was different in ASW, no 581 significant differences were found in terms of Au accumulation in the tissues of 582 S. aurata after the exposure to cAuNPs and PVP-AuNPs. Another interesting 583 584 result was the highest BAF in the spleen calculated for the lowest nominal concentration exposure of cAuNPs (4  $\mu$ g.L<sup>-1</sup>). This may be due to the lower 585 586 aggregation/agglomeration processes, with higher ability of Au entering in the tissues (Barreto et al. 2019a). As already reported, aggregation/agglomeration 587 is expected to increase with the increase in the number of particles per volume 588 (Barreto et al. 2015). Therefore, since cAuNPs, at the highest tested 589 concentrations (80 and 1600 µg.L<sup>-1</sup>), are more likely to aggregate/agglomerate, 590 they are less available for the uptake by fish. The detected accumulation of Au 591 in the spleen of S. aurata (although not significantly) show the potential role of 592 the spleen in Au elimination. Despite the lack of Au accumulation after the 593 exposure to AuNPs, particles were bioactive to S. aurata. The present data 594 reveal that 7 nm cAuNPs induced more pronounced effects, in terms of 595 oxidative stress and damage responses, than PVP-AuNPs. This result was 596

unexpected considering the stability of the particles, despite the fact that PVP 597 598 coating is considered safer and more biocompatible than citrate coating (Min et al. 2009). The 7 nm PVP-AuNPs remained stable in ASW, dispersed in the 599 water column and, therefore, more available for the uptake by fish. An opposite 600 pattern was previously observed for 40 nm AuNPs with PVP-AuNPs inducing 601 more adverse effects than cAuNPs to S. aurata (Barreto et al. 2019b). However, 602 603 a previous study, where different organisms were tested (bacteria, algae, SiHa cell line and mice), also showed that cAuNPs had more adverse effects than 604 PVP-AuNPs (Iswarya et al. 2016). Wang et al. (2011) reported that the toxicity 605 606 of AuNPs is related to the co-existence of citrate and Au<sup>3+</sup> ions. When NPs are 607 coated with PVP, there is an absence of reactive citrate ions on the surface of AuNPs. In the present study, the tested concentrations of cAuNPs decreased 608 the swimming resistance, induced enzymatic and non-enzymatic responses 609 involved in the oxidative defence/damage and genotoxicity in fish. Therefore, it 610 seems that the formed agglomerates/aggregates in the ASW (less than 200 nm) 611 may be incorporated through the cellular membranes. 612

613 Comparing the ionic with the nano form, Au significantly accumulated in 614 almost all assessed tissues of S. aurata after the exposure to 1600 µg.L<sup>-1</sup> iAu, whereas after the exposure to AuNPs, Au did not significantly accumulate. 615 Despite the significantly higher effects of cAuNPs in some endpoints (such as 616 617 gills GPx and liver GR activities, gills NPT and gills/liver LPO levels), even at the lowest tested concentration (i.e., 4 µg L<sup>-1</sup>), iAu induced, in general, more 618 effects on the fish (gills CAT, GR and liver GST activities and muscle/brain LPO 619 620 levels). Additionally, no fish mortality was detected after the exposures with AuNPs, whereas one fish died after the exposure to 1600 µg.L<sup>1</sup> of iAu. The 621

results demonstrated a tissue specificity, cAuNPs induced LPO in gills and liver, 622 while iAu induced LPO in brain and muscle. There are few available studies 623 about the mechanisms involved in the toxicity of iAu. Nonetheless, the iAu 624 625 ability to undergo redox reactions with peptides and proteins, particularly involving sulphur amino acids, to deprotonate and bind to peptide amide bonds 626 and cross-link histidine imidazole rings, has been already reported (Best and 627 Sadler 1996; Luis et al. 2016). Some authors, using in vitro tests, reported that 628 iAu induced effects to mussel (Mytilus galloprovincialis) and rainbow trout 629 630 (Oncorhynchus mykiss) whereas the nano form did not have any effect (Farkas et al. 2010; Luis et al. 2016). Botha el al. (2015), using different aquatic species 631 632 (daphnia and fish), also showed that iAu was more toxic than nano form. 633 However, Barbasz et al. (2016), using two types of human cell lines, showed a 634 higher cytotoxicity of AuNPs than iAu. Dedeh et al. (2014) described that, in spite of iAu having accumulated more in the tissues of zebrafish (Danio rerio) 635 than the AuNPs, the latter had more effects on the fish, in terms of gene 636 expression and neurotransmission. 637

In terms of genotoxicity, all the treatments induced DNA strand breaks, 638 assessed by comet assay, in S. aurata peripheral blood cells. Concerning 639 640 cytogenetic damage, ENAs frequency increased with the increase of Au concentration (nano or ionic form). A previous study also showed the potential 641 genotoxicity of 40 nm AuNPs (both PVP-AuNPs and cAuNPs) to S. aurata 642 (Barreto et al. 2019a). Comet assay is a rapid method to detect low levels of 643 DNA damage. However, this technique gives limited information about the kind 644 645 of DNA damage, if it is a direct consequence of the damaging agent or indirect effects, such as oxidative damage, apurinic/pyrimidinic sites or DNA repair 646

(Smith, O'Donovan, and Martin 2006). As previously described, the genotoxic 647 effects of AuNPs may be caused directly following the entry of NPs into the 648 nuclei, binding to DNA; or indirectly, through oxidative stress, which may 649 650 consequently induce DNA oxidative damage (Auffan et al. 2009; Barreto et al. 2019a; Cardoso et al. 2014). The production of reactive oxygen species (ROS) 651 following AuNPs exposure has been demonstrated in studies involving aquatic 652 organisms (Farkas et al. 2010; Tedesco et al. 2008, 2010; Pan et al. 2012). 653 654 Modification of the comet assay with the incorporation of lesion specific endonucleases, such as Fpg, increases its sensitivity and specificity through the 655 656 recognition of damaged bases and introduction of additional breaks (Azqueta et al. 2013; Smith, O'Donovan, and Martin 2006; Speit et al. 2004). The present 657 658 study showed that Au (nano and ionic form) induced DNA breaks, but oxidative 659 DNA damage was not observed. This result was previously described in studies with different types of NPs (Ag, CeO<sub>2</sub>, Co<sub>3</sub>O<sub>4</sub> and SiO<sub>2</sub>) and metal ions (Al<sup>3+</sup>, 660 Ni<sup>2+</sup>, Co<sup>2+</sup>, Cd<sup>2+</sup>, Cu<sup>2+</sup> and Zn<sup>2+</sup>) (Grin et al. 2009; Kain, Karlsson, and Möller 661 2012) and may be due to: 1) the low potential of the tested conditions to induce 662 oxidative damage on the erythrocyte DNA of S. aurata, which is not supported 663 664 by the LPO data; 2) the oxidative DNA lesions caused by the exposure to Au may have been already been repaired by cellular DNA repair systems (Catalán 665 et al. 2014); 3) NPs and ionic forms may interact with Fpg, not allowing the 666 binding of the enzyme with DNA (Asmuss et al. 2000; Kain, Karlsson, and 667 Möller 2012). 668

669 Overall, after Au exposures, enzymatic and non-enzymatic responses 670 involved in the defence of *S. aurata* against oxidative damage were more 671 activated in the gills than in the liver. Additionally, oxidative damage (LPO

increase) was more clearly expressed in gills than in liver. Gills are the first
organ to be exposed and provide a large surface area for contaminants such as
AuNPs, being considered a good candidate to an early assessment of the
effects of waterborne contaminants (Oliveira, Pacheco, and Santos 2008).

Since S. aurata is one of the most consumed fish in south Europe, an 676 estimation of Au intake by humans via food chain is an important assessment 677 (Barreto et al. 2019b). The highest estimated value for Au intake by each 678 Portuguese person (1.10 µg.kg body weight per year) would be relevant 679 following an exposure of the fish to 1600 µg.L<sup>-1</sup> iAu. Based on the tested 680 681 conditions and present results, the estimated maximum Au intake by humans per day was around 0.003 µg.kg<sup>-1</sup> body weight. So, this value did not exceed 682 the estimated TDI value for Au (322 µg.kg<sup>-1</sup>). Future studies should carry out the 683 684 assessment of Au intake by humans via food chain since AuNPs use is increasing worldwide and thus it is expected to find increased concentration of 685 686 them in the environment.

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# 688 **5. Conclusions**

The present results showed that short-term exposure to gold (nano or ionic 689 form), at low levels such as 4  $\mu$ g.L<sup>-1</sup>, was able to induce oxidative stress and 690 damage, as well as genotoxicity to the marine/estuarine fish Sparus aurata. 691 Citrate coated gold nanoparticles (cAuNPs), even aggregating/agglomerating in 692 seawater, induced significantly more effects to fish (oxidative stress and 693 damage) than the polyvinylpyrrolidone coated gold nanoparticles (PVP-AuNPs), 694 695 which maintained its nano size in seawater. The exposures to ionic gold resulted in higher accumulation in the fish tissues and induced more effects to 696

fish than nano form. After gold exposures, responses involved in fish defence against oxidative damage were more activated in the gills than in the liver. Furthermore, oxidative damage (lipid peroxidation increase) was more detected in gills than in the liver. The results showed that gold (nano and ionic form) is not inert and a distinct response was found in the assessed tissues. Further chronic tests must be performed to complement the present findings.

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- 704 **Conflict of interest statement**
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The authors declare that there are no conflicts of interest.

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#### 722 6. References

Albertini, R. J., D. Anderson, G. R. Douglas, L. Hagmar, K. Hemminki, F.
Merlo, A. T. Natarajan, H. Norppa, D. E. G. Shuker, R. Tice, Michael D. Waters,
and Antero Aitio. 2000. 'IPCS guidelines for the monitoring of genotoxic effects
of carcinogens in humans', *Mutation Research/Reviews in Mutation Research*,
463: 111-72.

Asmuss, M., L. H. F. Mullenders, A. Eker, and A. Hartwig. 2000. 'Differential effects of toxic metal compounds on the activities of Fpg and XPA, two zinc finger proteins involved in DNA repair', *Carcinogenesis*, 21: 2097-104.

Athar, M., and M. Iqbal. 1998. 'Ferric nitrilotriacetate promotes Ndiethylnitrosamine-induced renal tumorigenesis in the rat: implications for the involvement of oxidative stress', *Carcinogenesis*, 19: 1133-39.

Auffan, M., J. Rose, T. Orsiere, M. De Meo, A. Thill, O. Zeyons, O. Proux, A.
Masion, P. Chaurand, O. Spalla, A. Botta, M. R. Wiesner, and J.-Y. Bottero.
2009. 'CeO<sub>2</sub> nanoparticles induce DNA damage towards human dermal
fibroblasts *in vitro*', *Nanotoxicology*, 3: 161-71.

Azqueta, A., L. Arbillaga, A. L. de Cerain, and An. Collins. 2013. 'Enhancing the sensitivity of the comet assay as a genotoxicity test, by combining it with bacterial repair enzyme FPG', *Mutagenesis*, 28: 271-77.

Barbasz, A., and M. Oćwieja. 2016. 'Gold nanoparticles and ions – friends or
foes? As they are seen by human cells U-937 and HL-60', *Journal of Experimental Nanoscience*, 11: 564-80.

Barreto, A., L. G. Luis, A. M. V. M. Soares, P. Paíga, L. H. M. L. M. Santos,
C. Delerue-Matos, K. Hylland, S. Loureiro, and M. Oliveira. 2017. 'Genotoxicity

of gemfibrozil in the gilthead seabream (*Sparus aurata*)', *Mutation Research/Genetic Toxicology and Environmental Mutagenesis*, 821: 36-42.

Barreto, A., L. G. Luis, A. V. Girão, T. Trindade, A. M. V. M. Soares, and M.
Oliveira. 2015. 'Behavior of colloidal gold nanoparticles in different ionic
strength media', *Journal of Nanoparticle Research*, 17: 1-13.

Barreto, A., L. G. Luis, E. Pinto, A. Almeida, P. Paíga , L. H. M. L. M. Santos,
C. Delerue-Matos, T. Trindade, A. M. V. M. Soares, K. Hylland, S. Loureiro, M.
Oliveira. 2019a. 'Genotoxicity of gold nanoparticles in the gilthead seabream
(*Sparus aurata*) after single exposure and combined with the pharmaceutical
gemfibrozil', *Chemosphere*, 220:11-19.

Barreto, A., L. G. Luis, E. Pinto, A. Almeida, P. Paíga, L. H. M. L. M. Santos,
C. Delerue-Matos, T. Trindade, A. M. V. M. Soares, K. Hylland, S. Loureiro, M.
Oliveira. 2019b. 'Effects and bioaccumulation of gold nanoparticles in the
gilthead seabream (*Sparus aurata*) – Single and combined exposures with
gemfibrozil', *Chemosphere*, 215:248-60.

Best, S. L., and P. J. Sadler. 1996. 'Gold drugs: Mechanism of action and
toxicity', *Gold Bulletin*, 29: 87-93.

Bhattacharya, R., and P. Mukherjee. 2008. 'Biological properties of "naked"
 metal nanoparticles', *Advanced Drug Delivery Reviews*, 60: 1289-306.

Botha, T. L., T. E. James, and V. Wepener. 2015. 'Comparative Aquatic
Toxicity of Gold Nanoparticles and Ionic Gold Using a Species Sensitivity
Distribution Approach', *Journal of Nanomaterials*, 2015: 986902.

Boverhof, D. R., C. M. Bramante, J. H. Butala, S. F. Clancy, M. Lafranconi, J.
West, and S. C. Gordon. 2015. 'Comparative assessment of nanomaterial

definitions and safety evaluation considerations', *Regulatory Toxicology and Pharmacology*, 73: 137-50.

Bradford, M. M. 1976. 'A rapid and sensitive method for the quantitation of
microgram quantities of protein utilizing the principle of protein-dye binding', *Analytical Biochemistry*, 72: 248-54.

Bulloch, D. N., R. Lavado, K. L. Forsgren, S. Beni, D. Schlenk, and C. K.
Larive. 2012. 'Analytical and Biological Characterization of Halogenated
Gemfibrozil Produced through Chlorination of Wastewater', *Environmental Science & Technology*, 46: 5583-89.

Cardoso, E., G. T. Rezin, E. T. Zanoni, F. de S. Notoya, D. D. Leffa, A. P.
Damiani, F. Daumann, J. C. O. Rodriguez, R. Benavides, L. da Silva, V. M.
Andrade, and P. M. M. da Silva. 2014. 'Acute and chronic administration of gold
nanoparticles cause DNA damage in the cerebral cortex of adult rats', *Mutation research*, 766-767: 25-30.

Carlberg, I., and B Mannervik. 1975. 'Purification and characterization of the flavoenzyme glutathione reductase from rat liver', *Journal of Biological Chemistry*, 250: 5475-80.

Catalán, J., S. Suhonen, A. Huk, and M. Dusinska. 2014. 'Analysis of
nanoparticle-induced DNA damage by the comet assay.' in L. María Sierra and
Isabel Gaivão (eds.), *Genotoxicity and DNA Repair: A Practical Approach*(Springer New York: New York, NY).

Claiborne, A. 1985. 'Catalase activity', *CRC handbook of methods for oxygen radical research*, 1: 283-84.

Collins, A. R. 2014. 'Measuring oxidative damage to DNA and its repair with
the comet assay', *Biochimica et Biophysica Acta (BBA) - General Subjects*,
1840: 794-800.

Collins, A. R., V. L. Dobson, M. Dušinská, G. Kennedy, and R. Štětina. 1997.
'The comet assay: what can it really tell us?', *Mutation Research/Fundamental*

798 and Molecular Mechanisms of Mutagenesis, 375: 183-93.

Collins, A. R. 2004. 'The comet assay for DNA damage and repair', *Molecular Biotechnology*, 26: 249-61.

Coradeghini, R., S. Gioria, C. P. García, P. Nativo, F. Franchini, D. Gilliland, J. Ponti, and F. Rossi. 2013. 'Size-dependent toxicity and cell interaction mechanisms of gold nanoparticles on mouse fibroblasts', *Toxicology Letters*, 217: 205-16.

Daniel, M.-C., and D. Astruc. 2004. 'Gold nanoparticles: assembly, supramolecular chemistry, quantum-size-related properties, and applications toward biology, catalysis, and nanotechnology', *Chemical Reviews*, 104: 293-346.

Das, T., V. Kolli, S. Karmakar, and N. Sarkar. 2017. 'Functionalisation of polyvinylpyrrolidone on gold nanoparticles enhances its anti-amyloidogenic propensity towards hen egg white lysozyme', *Biomedicines*, 5: 19.

de Andrade, V. M., T. R. O. de Freitas, and J. da Silva. 2004. 'Comet assay using mullet (*Mugil* sp.) and sea catfish (*Netuma* sp.) erythrocytes for the detection of genotoxic pollutants in aquatic environment', *Mutation Research/Genetic Toxicology and Environmental Mutagenesis*, 560: 57-67.

Dedeh, A., A. Ciutat, M. Treguer-Delapierre, and J.-P. Bourdineaud. 2015. 'Impact of gold nanoparticles on zebrafish exposed to a spiked sediment', *Nanotoxicology*, 9: 71-80.

Ellman, G. L., K. D. Courtney, V. Andres jr, and R. M. Featherstone. 1961. 'A new and rapid colorimetric determination of acetylcholinesterase activity', *Biochemical Pharmacology*, 7: 88-95.

Epe, B., M. Pflaum, M. Häring, J. Hegler, and H. Rüdiger. 1993. 'Use of repair endonucleases to characterize DNA damage induced by reactive oxygen species in cellular and cell-free systems', *Toxicology Letters*, 67: 57-72.

Failler, P., G. V. de Walle, N. Lecrivain, A. Himbes, and R. Lewins. 2007. Future prospects for fish and fishery products. 4. Fish consumption in the European Union in 2015 and 2030. Part 1', *European overview FAO Fisheries Circular (FAO)*.

Farkas, J., P. Christian, J. A. G. Urrea, N. Roos, M. Hassellöv, K. E.
Tollefsen, and K. V. Thomas. 2010. 'Effects of silver and gold nanoparticles on
rainbow trout (*Oncorhynchus mykiss*) hepatocytes', *Aquatic Toxicology*, 96: 4452.

FDA. 2015. 'Q3D Elemental Impurities Guidance for Industry'.

Filho, D., T. Tribess, C. Gáspari, F. D. Claudio, M. A. Torres, and A. R. M.
Magalhães. 2001. 'Seasonal changes in antioxidant defenses of the digestive
gland of the brown mussel (*Perna perna*)', *Aquaculture*, 203: 149-58.

Frasco, M. F., and L. Guilhermino. 2002. 'Effects of dimethoate and betanaphthoflavone on selected biomarkers of *Poecilia reticulata*', *Fish Physiology and Biochemistry*, 26: 149-56.

840 Fratoddi, I., I. Venditti, C. Cametti, and M. V. Russo. 2015. 'How toxic are 841 gold nanoparticles? The state-of-the-art', *Nano Research*, 8: 1771-99.

García-Negrete, C. A., J. Blasco, M. Volland, T. C. Rojas, M. Hampel, A.
Lapresta-Fernández, M. C. Jiménez de Haro, M. Soto, and A. Fernández. 2013.
'Behaviour of Au-citrate nanoparticles in seawater and accumulation in bivalves
at environmentally relevant concentrations', *Environmental Pollution*, 174: 13441.

Ghosh, P., G. Han, M. De, C. K. Kim, and V. M. Rotello. 2008. 'Gold nanoparticles in delivery applications', *Advanced Drug Delivery Reviews*, 60: 1307-15.

Grin, I., P. G Konorovsky, G. Nevinsky, and D. O Zharkov. 2009. 'Heavy metal ions affect the activity of DNA glycosylases of the Fpg family', *Biochemistry*, 74: 1253-9.

Guilhermino, L., M. C. Lopes, A. P. Carvalho, and A. M. V. M. Soares. 1996.
'Acetylcholinesterase activity in juveniles of *Daphnia magna* straus', *Bulletin of Environmental Contamination and Toxicology*, 57: 979-85.

Habig, W. H., Michael J. Pabst, and W. B. Jakoby. 1974. 'Glutathione STransferases. The first enzymatic step in mercapturic acid formation', *Journal of Biological Chemistry*, 249: 7130-39.

Hanžić, N., T. Jurkin, A. Maksimović, and M. Gotić. 2015. 'The synthesis of
gold nanoparticles by a citrate-radiolytical method', *Radiation Physics and Chemistry*, 106: 77-82.

Ido, A., Y. Hiromori, L. Meng, H. Usuda, H. Nagase, M. Yang, J. Hu, and T.
Nakanishi. 2017. 'Occurrence of fibrates and their metabolites in source and
drinking water in Shanghai and Zhejiang, China', *Scientific Reports*, 7: 45931.

865 IPCS. 2004. 'Harmonization project document no. 1 - IPCS risk assessment 866 terminology', *WHO, Geneva*.

Iswarya, V., J. Manivannan, A. De, S. Paul, R. Roy, J. B. Johnson, R. Kundu,
N. Chandrasekaran, A. Mukherjee, and A. Mukherjee. 2016. 'Surface capping
and size-dependent toxicity of gold nanoparticles on different trophic levels', *Environmental Science and Pollution Research*, 23: 4844-58.

Jackevicius, C. A., J. V. Tu, J. S. Ross, D. T. Ko, D. Carreon, and H. M. Krumholz. 2011. 'Use of Fibrates in the United States and Canada', *Jama*, 305: 1217-24.

Kain, J., H. L. Karlsson, and L. Möller. 2012. 'DNA damage induced by microand nanoparticles-interaction with FPG influences the detection of DNA
oxidation in the comet assay', *Mutagenesis*, 27: 491-500.

Kim, K., H. K. Kleinman, H.-J. Lee, and K. Pahan. 2017. 'Safety and potential
efficacy of gemfibrozil as a supportive treatment for children with late infantile
neuronal ceroid lipofuscinosis and other lipid storage disorders', *Orphanet Journal of Rare Diseases*, 12: 113.

Li, C., D. Li, G. Wan, J. Xu, and W. Hou. 2011. 'Facile synthesis of concentrated gold nanoparticles with low size-distribution in water: temperature and pH controls', *Nanoscale Research Letters*, 6: 1-10.

Li, Q., J. Laval, and D. B. Ludlum. 1997. 'Fpg protein releases a ring-opened N-7 guanine adduct from DNA that has been modified by sulfur mustard', *Carcinogenesis*, 18:1035-38.

Lima, I., S. M. Moreira, J. R.-V. Osten, A. M. V. M. Soares, and L. Guilhermino. 2007. 'Biochemical responses of the marine mussel *Mytilus* 

*galloprovincialis* to petrochemical environmental contamination along the Northwestern coast of Portugal', *Chemosphere*, 66: 1230-42.

Liu, X., M. Atwater, J. Wang, and Q. Huo. 2007. 'Extinction coefficient of gold nanoparticles with different sizes and different capping ligands', *Colloids and Surfaces B: Biointerfaces*, 58: 3-7.

Loganathan, C., and S. A. John. 2017. 'Naked eye and spectrophotometric detection of chromogenic insecticide in aquaculture using amine functionalized gold nanoparticles in the presence of major interferents', *Spectrochimica Acta Part A: Molecular and Biomolecular Spectroscopy*, 173: 837-42.

Luis, L. G., A. Barreto, T. Trindade, A. M. V. M. Soares, and M. Oliveira. 2016. 'Effects of emerging contaminants on neurotransmission and biotransformation in marine organisms – An *in vitro* approach', *Marine Pollution Bulletin*, 106: 236-44.

Mahl, D., C. Greulich, W. Meyer-Zaika, M. Koller, and M. Epple. 2010. 'Gold
nanoparticles: dispersibility in biological media and cell-biological effect', *Journal of Materials Chemistry*, 20: 6176-81.

Min, Z., W. Baoxiang, R. Zbigniew, X. Zhaohui, F. J. Otto, Y. Xiaofeng, and
R. Steinar. 2009. 'Minute synthesis of extremely stable gold nanoparticles', *Nanotechnology*, 20: 505606.

Mohandas, J., J. J. Marshall, G. G. Duggin, J. S. Horvath, and D. J. Tiller. 1984. 'Differential distribution of glutathione and glutathione-related enzymes in rabbit kidney', *Biochemical Pharmacology*, 33: 1801-07.

Na-Phatthalung, P., M. Teles, L. Tort, and M. Oliveira. 2018. 'Gold nanoparticles exposure modulates antioxidant and innate immune gene expression in the gills of *Sparus aurata*', *Genomics*, 110: 430-34.

914 NIST. 2010. 'NCL Method PCC-8, Determination of gold in rat tissue with 915 inductively coupled plasma mass spectrometry'.

Nunes, B., F. Carvalho, and L. Guilhermino. 2004. 'Acute and chronic effects
of clofibrate and clofibric acid on the enzymes acetylcholinesterase, lactate
dehydrogenase and catalase of the mosquitofish, *Gambusia holbrooki*', *Chemosphere*, 57: 1581-89.

920 OECD. 1992. Test No. 203: Fish, Acute Toxicity Test (OECD Publishing).

Ohkawa, H., N. Ohishi, and K. Yagi. 1979. 'Assay for lipid peroxides in animal tissues by thiobarbituric acid reaction', *Analytical Biochemistry*, 95: 351-58.

Oliveira, M., C. Gravato, and L. Guilhermino. 2012. 'Acute toxic effects of pyrene on *Pomatoschistus microps* (Teleostei, Gobiidae): Mortality, biomarkers and swimming performance', *Ecological Indicators*, 19: 206-14.

Oliveira, M., M. Pacheco, and M. A. Santos. 2008. 'Organ specific antioxidant responses in golden grey mullet (*Liza aurata*) following a short-term exposure to phenanthrene', *Science of The Total Environment*, 396: 70-78.

Ozansoy, G., B. Akin, F. Aktan, and Ç. Karasu. 2001. 'Short-term gemfibrozil treatment reverses lipid profile and peroxidation but does not alter blood glucose and tissue antioxidant enzymes in chronically diabetic rats', *Molecular and Cellular Biochemistry*, 216: 59-63.

Pacheco, M., and M. A. Santos. 1996. 'Induction of micronuclei and nuclear abnormalities in the erythrocytes of *Anguilla anguilla* L. exposed either to cyclophosphamide or to bleached kraft pulp mill efluent', *Fresenius Environmental Bulletin*, 5: 746-51.

Pan, J.-F., P.-E. Buffet, L. Poirier, C. Amiard-Triquet, D. Gilliland, Y. Joubert,
P. Pilet, M. Guibbolini, C. R. de Faverney, M. Roméo, E. Valsami-Jones, and C.
Mouneyrac. 2012. 'Size dependent bioaccumulation and ecotoxicity of gold
nanoparticles in an endobenthic invertebrate: The Tellinid clam *Scrobicularia plana*', *Environmental Pollution*, 168: 37-43.

Paramelle, D., A. Sadovoy, S. Gorelik, P. Free, J. Hobley, and D. G. Fernig.
2014. 'A rapid method to estimate the concentration of citrate capped silver
nanoparticles from UV-visible light spectra', *Analyst*, 139: 4855-61.

Parvez, S., I. Sayeed, S. Pandey, A. Ahmad, B. Bin-Hafeez, R. Haque,
I.Ahmad, and S. Raisuddin. 2003. 'Modulatory effect of copper on nonenzymatic
antioxidants in freshwater fish *Channa punctatus* (Bloch.)', *Biological Trace Element Research*, 93: 237-48.

Prindiville, J. S., J. A. Mennigen, J. M. Zamora, T. W. Moon, and J.-M.
Weber. 2011. 'The fibrate drug gemfibrozil disrupts lipoprotein metabolism in
rainbow trout', *Toxicology and Applied Pharmacology*, 251: 201-08.

Qu, B., Q.-T. Li, K. P. Wong, T. M. C. Tan, and B. Halliwell. 2001.
'Mechanism of clofibrate hepatotoxicity: mitochondrial damage and oxidative
stress in hepatocytes', *Free Radical Biology and Medicine*, 31: 659-69.

Quinn, B., W. Schmidt, K. O'Rourke, and R. Hernan. 2011. 'Effects of the pharmaceuticals gemfibrozil and diclofenac on biomarker expression in the zebra mussel (*Dreissena polymorpha*) and their comparison with standardised toxicity tests', *Chemosphere*, 84: 657-63.

Roy, A., and K. Pahan. 2009. 'Gemfibrozil, stretching arms beyond lipid
lowering', *Immunopharmacology and immunotoxicology*, 31: 339-51.

- Saleh, M., G. Kumar, A.-A. Abdel-Baki, S. Al-Quraishy, and M. El-Matbouli.
  2016. '*In vitro* antimicrosporidial activity of gold nanoparticles against *Heterosporis saurida*', *BMC Veterinary Research*, 12: 44.
- Sedlak, J., and R. H. Lindsay. 1968. 'Estimation of total, protein-bound, and
  nonprotein sulfhydryl groups in tissue with Ellman's reagent', *Analytical Biochemistry*, 25: 192-205.
- Shiba, F. 2013. 'Size control of monodisperse Au nanoparticles synthesized
  via a citrate reduction process associated with a pH-shifting procedure', *CrystEngComm*, 15: 8412-15.
- 971 SigmaPlot 12.0. 2011. 'Statistical Package for the Social Sciences 972 SigmaPlot for Windows', 12.0 ed, Chicago, IL, USA.
- Singh, N. P., M. T. McCoy, R. R. Tice, and E. L. Schneider. 1988. 'A simple
  technique for quantitation of low levels of DNA damage in individual cells', *Experimental Cell Research*, 175: 184-91.
- 976 Skolness, S. Y., E. J. Durhan, K. M. Jensen, M. D. Kahl, E. A. Makynen, D.L.
- 977 Villeneuve, and G. T. Ankley. 2012. 'Effects of gemfibrozil on lipid metabolism,
- 978 steroidogenesis, and reproduction in the fathead minnow (Pimephales
- promelas)', Environmental Toxicology and Chemistry, 31: 2615-24.
- Smith, C. C., M. R. O'Donovan, and E. A. Martin. 2006. 'hOGG1 recognizes
  oxidative damage using the comet assay with greater specificity than FPG or
  ENDOIII', *Mutagenesis*, 21: 185-90.
- Speit, G., P. Schütz, I. Bonzheim, K. Trenz, and H. Hoffmann. 2004.
  'Sensitivity of the FPG protein towards alkylation damage in the comet assay', *Toxicology Letters*, 146: 151-58.

Sutken, E., M. Inal, and F. Ozdemir. 2006. 'Effects of vitamin E and gemfibrozil on lipid profiles, lipid peroxidation and antioxidant status in the elderly and young hyperlipidemic subjects', *Saudi Medical Journal*, 27: 453-59.

Tchou, J., V. Bodepudi, S. Shibutani, I. Antoshechkin, J. Miller, A. P.
Grollman, and F. Johnson. 1994. 'Substrate specificity of Fpg protein.
Recognition and cleavage of oxidatively damaged DNA', *Journal of Biological Chemistry*, 269: 15318-24.

Tedesco, S., H. Doyle, G. Redmond, and D. Sheehan. 2008. 'Gold
nanoparticles and oxidative stress in *Mytilus edulis*', *Marine Environmental Research*, 66: 131-33.

Tedesco, S., H. Doyle, J. Blasco, G. Redmond, and D. Sheehan. 2010.
'Oxidative stress and toxicity of gold nanoparticles in *Mytilus edulis*', *Aquatic Toxicology*, 100: 178-86.

Tejamaya, M., I. Römer, R. C. Merrifield, and J. R. Lead. 2012. 'Stability of
citrate, PVP, and PEG coated silver nanoparticles in ecotoxicology media', *Environmental Science & Technology*, 46: 7011-17.

Teles, M., C. Fierro-Castro, P. Na-Phatthalung, A. Tvarijonaviciute, T. Trindade, A. M. V. M. Soares, L. Tort, and M. Oliveira. 2016. 'Assessment of gold nanoparticle effects in a marine teleost (*Sparus aurata*) using molecular and biochemical biomarkers', *Aquatic Toxicology*, 177: 125-35.

Termini, J. 2000. 'Hydroperoxide-induced DNA damage and mutations', *Mutation Research - Fundamental and Molecular Mechanisms of Mutagenesis*,
450: 107-24.

1009 Tice, R. R., E. Agurell, D. Anderson, B. Burlinson, A. Hartmann, H. 1010 Kobayashi, Y. Miyamae, E. Rojas, J. C. Ryu, and Y. F. Sasaki. 2000. 'Single

1011 cell gel/comet assay: Guidelines for *in vitro* and *in vivo* genetic toxicology 1012 testing', *Environmental and Molecular Mutagenesis*, 35: 206-21.

1013 Tiede, K., M. Hassellöv, E. Breitbarth, Q. Chaudhry, and A. B. A. Boxall. 1014 2009. 'Considerations for environmental fate and ecotoxicity testing to support 1015 environmental risk assessments for engineered nanoparticles', *Journal of* 1016 *Chromatography A*, 1216: 503-09.

Tudek, B., A. A. VanZeeland, J. T. Kusmierek, and J. Laval. 1998. 'Activity of *Escherichia coli* DNA-glycosylases on DNA damaged by methylating and ethylating agents and influence of 3-substituted adenine derivatives', *Mutation Research/DNA Repair*, 407: 169-76.

1021 Turkevich, J., P. C. Stevenson, and J. Hillier. 1951. 'A study of the nucleation 1022 and growth processes in the synthesis of colloidal gold', *Discussions of the* 1023 *Faraday Society*, 11: 55-75.

Vale, G., K. Mehennaoui, S. Cambier, G. Libralato, S. Jomini, and R. F. Domingos. 2016. 'Manufactured nanoparticles in the aquatic environmentbiochemical responses on freshwater organisms: A critical overview', *Aquatic Toxicology*, 170: 162-74.

Vieira, H. C., F. Morgado, A. M. V. M. Soares, and S. N. Abreu. 2015. 'Fish consumption recommendations to conform to current advice in regard to mercury intake', *Environmental Science and Pollution Research*, 22: 9595-602.

1031 Wang, S., R. Lawson, P.C. Ray, H. Yu. 2011. 'Toxic effects of gold 1032 nanoparticles on *Salmonella typhimurium* bacteria', *Toxicology and Industrial* 1033 *Health*, 27: 547-54.

1034 WHO. 2008. 'Guidance for identifying populations at risk from mercury 1035 exposure', *Switzerland, Geneva*.

1036 Xia, Q., H. Li, Y. Liu, S. Zhang, Q. Feng, and K. Xiao. 2017. 'The effect of 1037 particle size on the genotoxicity of gold nanoparticles', *Journal of Biomedical* 1038 *Materials Research Part A*, 105: 710-19.

Yoo-lam, M., R. Chaichana, and T. Satapanajaru. 2014. 'Toxicity, bioaccumulation and biomagnification of silver nanoparticles in green algae (*Chlorella* sp.), water flea (*Moina macrocopa*), blood worm (*Chironomus* spp.) and silver barb (*Barbonymus gonionotus*)', *Chemical Speciation and Bioavailability*, 26: 257-65.

Zeng, S., M. Cai, H. Liang, and J. Hao. 2012. 'Size-dependent colorimetric
visual detection of melamine in milk at 10 ppb level by citrate-stabilized Au
nanoparticles', *Analytical Methods*, 4: 2499-505.